

1
2
3
4 **Metagenomics-based evaluation of groundwater microbial profiles in response to treated**
5
6
7 **wastewater discharge**
8
9

10
11
12
13
14
15 Noor Zaouri ¹, Muhammad Raihan Jumat ², Tariq Cheema ³ and Pei-Ying Hong ^{1*}
16
17
18
19
20
21
22
23
24

25 ¹ Water Desalination and Reuse Center (WDRC), Biological and Environmental Sciences and
26 Engineering Division (BESE), King Abdullah University of Science and Technology (KAUST),
27
28 Thuwal-Jeddah 23955-6900, Saudi Arabia
29
30
31

32 ² Duke-NUS Graduate Medical School, 8 College Road, Singapore 169857
33
34

35 ³ Hydrology Department, Faculty of Earth Sciences, King Abdulaziz University (KAU), Jeddah,
36
37 Saudi Arabia
38
39
40
41
42
43

44 * Corresponding author:
45
46

47 Pei-Ying Hong
48

49 Tel: +966-(12)-808-2218;
50
51

52 Email: peiying.hong@kaust.edu.sa
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Highlights

- Metagenomics as a tool to assess groundwater quality
- Lower alpha diversity of bacteria and virus due to treated wastewater exposure
- Increase in relative abundance of Planctomycetes and Picornaviridae
- Antibiotic resistance genes may not be good markers to differentiate wastewater exposure

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

2 **Abstract**

3 This study aims to demonstrate the use of metagenomics to assess groundwater quality.
4 Metagenomics revealed a lower alpha diversity for both bacteria and virus in wastewater-
5 exposed groundwater compared to the upstream controls. However, an increase in the relative
6 abundance of Planctomycetes and Picornaviridae, as well as increasing counts of unique
7 antibiotic resistance genes, was observed in wastewater-exposed groundwater. Findings suggest
8 that metagenomics can detect microbial signatures indicative of treated wastewater discharge.

9
10
11
12

13 **Keywords:** Shotgun metagenomics; aquifer recharge; groundwater quality; water reuse

1. Introduction

Water scarcity is projected to exacerbate by 2040, with 33 countries including Saudi Arabia predicted to face extreme water scarcity by that time (Luo et al., 2015). In recent years, recharging aquifer with treated wastewater has been proposed as a mitigation strategy to replenish the groundwater supplies. In Saudi Arabia, most wastewater treatment plants (WWTPs) are designed to achieve secondary or tertiary treatment with a final chlorination step prior to discharge. Such WWTPs typically achieve a cumulative 6-log reduction of microorganisms (Al-Jassim et al., 2015). Considering that untreated wastewater generally has a cell density of 10^{11} to 10^{12} cells/L (Manti et al., 2008), the final treated wastewater still contains a significant microbial load that may potentially perturb the indigenous groundwater microbial community. Aquifers are sensitive ecosystems, and therefore recharging aquifer with treated wastewater can only be adopted with an improved understanding on how these treated wastewater would affect the existing groundwater.

The current approaches to detecting changes in microbial community include qualitative approaches like terminal restriction fragment length polymorphism (T-RFLP) or quantitative approaches like real time PCR (qPCR). T-RFLP generates fingerprinting profiles for comparative analysis but does not provide information on the taxonomic identities of microbial contaminants. qPCR is a targeted approach that does not provide a comprehensive outlook on the wide variety of bacteria, virus and genes contaminants present in treated wastewater. The advent of high throughput untargeted sequencing approaches (e.g. metagenomics) have allowed a simultaneous evaluation on the bacteria, virus and genes present in an environmental sample (Quince et al., 2017). By comparing the diversity and relative abundance of these biological contaminants, one can denote if the groundwater has been perturbed by the treated wastewater.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

This study aims to demonstrate the use of metagenomics as a tool to assess microbial water quality. Changes in the diversity and relative abundance of bacteria, virus and antibiotic resistance genes were monitored to denote the extent of perturbation imposed by treated wastewater.

2. Material and method

2.1. Description of sampling sites and sampling procedure

A subsurface reservoir, otherwise also known as Wadi Fatimah in Arabic, situated along the Yamaniyah-Az Zibarah-Fatimah drainage system in Saudi Arabia (Grainger, 1992) was identified for groundwater monitoring. Wadi Fatimah has been receiving treated wastewaters as part of a study commissioned in 2010-2016 by Makkah water utilities to recharge local aquifers. Groundwater at this site was collected over a period of 9 months, and include the following: wells 1, 15, 14 and 17 in March 2016; wells 1, 14 and 17 in August 2016; wells 12, 15 and 17 in December 2016 (Figure S1). Treated wastewaters injected to Wadi Fatimah were from a local WWTP upstream of well 1, and utilized schematics including clarifiers, activated sludge tanks, oxidation ponds and chlorine disinfection. Sampling wells varied throughout the sampling period because of various reasons including lack of sufficient water to facilitate representative sampling, groundwater well was closed for access or no permission was granted from the well owners on the day of sampling. Two wells situated approximately 60 km upstream of Wadi Fatimah were also identified, and groundwater was collected from these wells during the same time frame as mentioned above. Samples from these two wells serve as upstream controls since both the wells are not known to actively receive any treated wastewater discharge (Figure S1). The collected volume of all groundwater samples ranged from 20-40 L per well, and were

1
2
3
4 61 comprised of a single grab sample collected per well per sampling event. Before collection, the
5
6 62 water was flushed for > 5 min to remove stagnant pools of water retained within the sampling
7
8
9 63 pipes. Water was also flushed until no apparent turbidity was observed prior to collection in
10
11 64 sterile polyethylene bottles. All samples were placed into a cooler and kept cool during transport.
12
13
14 65 Upon arrival in laboratory, samples were stored in 4 °C prior to analyses.
15
16

17 66 **2.2. Groundwater concentration and DNA extraction**

18
19
20 67 Each 20-40 L of groundwater sample was passed through a filtration system fitted with a T-
21
22 68 series cassette with a 100-kDa cut-off (approximately 0.005 µm) (Pall Corporation, Port
23
24
25 69 Washington, NY, USA). This cassette has been shown previously to retain viruses at an
26
27
28 70 efficiency of up to 87% and was particularly suited for large volume concentration (Jumat et al.,
29
30 71 2017). Hence, this procedure was chosen to concentrate any microbial matter present in the low
31
32
33 72 turbidity groundwater samples such as that collected in this study. Each groundwater sample was
34
35 73 concentrated to 5 mL of retentate. To collect any microbial matter adhering to the 100 kDa
36
37
38 74 cassette, the system was washed with 50 mL of wash solution (1 x Phosphate Buffered Saline
39
40 75 (PBS) and 0.01% Tween 60). After circulating the wash solution through the filtration system for
41
42
43 76 several minutes, the wash solution was itself concentrated to 5 mL. Both retentate and wash
44
45 77 solutions derived from the same sample were extracted for DNA was by using the DNeasy Blood
46
47 78 and Tissue Kit (Qiagen, Hilgen, Germany). MilliQ water was collected and subjected to the same
48
49
50 79 filtration procedure at the same time when all groundwater samples were filtered. This acts as a
51
52 80 negative control, and provide comparative analysis that would allow determination of microbial
53
54
55 81 populations unique to the groundwater.
56

57 82 **2.3. Metagenomic sequencing and analysis**

1
2
3
4 83 Genomic DNA was sent to CosmosID (Rockville, MD, USA) for sequencing on HiSeq, as
5
6 84 described (Hasan et al., 2014). Shotgun metagenomic datasets were directly analyzed by
7
8 85 CosmosID metagenomic software (CosmosID Inc., Rockville, MD) as described elsewhere
9
10 86 (Connelly et al., 2018; Hasan et al., 2014; Kaleko et al., 2016; Ponnusamy et al., 2016) to reveal
11
12 87 associated microbial community composition. Briefly, the system utilizes a high performance
13
14 88 data-mining k-mer algorithm and highly curated dynamic comparator databases (GenBook®)
15
16 89 that rapidly disambiguate millions of short reads into the discrete genomes or genes engendering
17
18 90 the particular sequences. The GeneBook databases were composed of over 150,000 microbial
19
20 91 genomes and gene sequences representing over 1000 bacterial and 5000 viral species, as well as
21
22 92 over 5500 antibiotic resistance genes. For each sample, the reads from a bacteria or viral species
23
24 93 are assigned to the strain with the highest aggregation statistics, and subsequently determined for
25
26 94 the corresponding alpha diversity. Community resistome was also identified using the CosmosID
27
28 95 Inc. bioinformatics software package by querying the unassembled sequence reads against the
29
30 96 CosmosID curated antibiotic resistance gene databases. Paired end Fastq files are deposited onto
31
32 97 European Nucleotide Archive under accession number PRJEB32173.
33
34
35
36
37
38
39
40

41 98 **2.4. Determination of organic water quality**

42
43
44 99 Organic constituents of the DOC were fractionated according to molecular weight (MW) with
45
46 100 size exclusion chromatography. The fraction distribution of the DOC in the groundwater was
47
48 101 analyzed using Liquid Chromatography equipped with a macro-porous silica separation column,
49
50 102 an Organic Carbon Detector (OCD), a UV 254 nm Detector (UVD) and an Organic Nitrogen
51
52 103 Detector (OND) (LC-OCD Model 8, DOC LABOR, Germany). ChromCaluni software was used
53
54 104 to determine the concentration of each fraction by integrating the area underneath the curve
55
56 105 according to manufacture instructions.
57
58
59
60
61
62
63
64
65

106 3. Results and discussion

107 3.1. Taxonomic profile of bacterial communities

108
109 Phyla Proteobacteria, Actinobacteria and Bacteroidetes constitute the three dominant bacterial
110 groups in the groundwater samples of both Wadi Fatimah and upstream controls. However,
111 Proteobacteria and Actinobacteria were also present in high relative abundance in the negative
112 MilliQ controls (Figure 1a). In contrast, Bacteroidetes was detected in higher relative abundance
113 and in more consistent frequency in only the groundwater samples. Planctomycetes was detected
114 in higher relative abundance among the Wadi Fatimah samples than that in upstream and
115 negative controls (Figure 1a). Planctomycetes has been shown to be closely associated with
116 organic detritus in the marine environment (Halliday et al., 2014). Similarly, a higher relative
117 abundance of Planctomycetes was observed in groundwater samples that had a higher dissolved
118 organic carbon (Table 1). Among these five detected DOC fractions, building block (BB) and
119 humic substances (HS) were significantly higher in groundwater sampled from Wadi Fatimah
120 compared to upstream controls ($p < 0.10$, Table 1). Specifically, HS which is predominated by
121 aromatic moieties was only detected in Wadi Fatimah at concentrations ranging from 0.54 to
122 0.70 mg/L. Despite the higher DOC concentrations in Wadi Fatimah, the alpha diversity of
123 bacterial community was lower in all of its samples compared to the controls (Figure 1b),
124 suggesting the presence of an unknown factor that bias against certain families (e.g.
125 Bradyrhizobiaceae, Methylobacteriaceae) in Wadi Fatimah. These families were only detected in
126 the upstream controls during the sampling period. There was however, a positive trend of
127 increasing alpha diversity with time, although the reason behind this trend remains unknown.

128 3.2. Community resistome

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

129 Ninety six unique ARGs were detected in Wadi Fatimah while only 56 unique ARGs were
130 detected in the upstream controls (Figure 1c). The 5 most abundant unique ARGs detected in
131 Wadi Fatimah were related to aminoglycoside, beta-lactam, macrolide and quinolone resistance,
132 while the 5 most abundant ARGs detected in upstream controls also included resistance against
133 beta-lactam. Shared ARGs between Wadi Fatimah and control samples revealed additional genes
134 conferring resistance to tetracycline and trimethoprim (Table 2). A positive detection of ARGs in
135 both Wadi Fatimah and control samples suggests a ubiquitous presence of ARGs in most
136 environments, including pristine environment (Van Goethem et al., 2018). This may be related to
137 long term global consumption of antibiotics for both hospital-associated and community-
138 acquired infections (Versporten et al., 2018). Hence, the choice of ARGs as metagenomics
139 markers to denote anthropogenic contamination events may not be provide definitive results.

3.3. Taxonomic profile of viral communities

141 Podoviridae and Siphoviridae account as the two most dominant viral families in the
142 groundwater (Figure 2a). It was further observed that upstream controls had a significantly
143 higher relative abundance of Siphoviridae, along with Myoviridae, Inoviridae and Tectiviridae (p
144 < 0.10). All four families are mainly comprised of bacteriophages that rely on bacteria and
145 archaea as their natural hosts. Along with one of the negative controls which denoted an
146 exceedingly high relative abundance of Picornaviridae, Picornaviridae was consistently detected
147 in all groundwater samples and especially at a higher relative abundance in Wadi Fatimah
148 samples (average 45%) than in the upstream controls (average 11%). Picornaviridae comprises
149 of several notable human enteric viruses including Enteroviruses and Salivirus which are
150 proposed as potential human virus indicator (Reuter et al., 2017). The high relative abundance of
151 Picornaviridae may be related to the type of treatment process used by the WWTP situated in

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

152 Wadi Fatimah. Secondary and tertiary treatment processes typically do not achieve high removal
153 rates of viruses (Zhang and Farahbakhsh, 2007). In contrast, a WWTP retrofitted with membrane
154 bioreactor reported only sporadic occurrences of Enterovirus (Jumat et al., 2017). The alpha
155 diversity of viruses are higher in the controls than in Wadi Fatimah (Figure 2b). Coupled with
156 the higher abundance of families associated with bacteriophages, the higher alpha diversity may
157 be due to a wide variety of bacteriophages present in relatively unperturbed environments
158 compared to those exposed to treated wastewater. Unlike the bacterial alpha diversity, the viruses
159 did not exhibit an increasing trend of alpha diversity with time (Figure 2b).

160

161 4. Conclusion

162 Findings from this study demonstrates the use of metagenomics to provide a simultaneous
163 evaluation of bacteria, virus and antibiotic resistome in groundwater exposed to treated
164 wastewater discharge. It was observed that alpha diversity of both bacteria and viruses were
165 lower in exposed groundwater than the upstream controls. In addition, Planctomycetes and
166 Picornaviridae, both of which were detected by metagenomics simultaneously, may be
167 appropriate markers to indicate increased organic content and presence of human fecal
168 contamination. As more countries explore the use of treated wastewater to recharge diminishing
169 groundwater levels, metagenomics could suffice as an appropriate monitoring tool to ensure that
170 a wide diversity of biological contaminants are assessed and unintentional aberrations to the
171 groundwater minimized. However, metagenomics alone is unable to provide insights to the
172 potential risks that may arise from recharging aquifers with treated wastewater. This is because
173 metagenomics analysis does not provide absolute quantification of biological contaminants and
174 are derived from DNA template sequencing. Hence it is unable to differentiate the viable bacteria

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

175 populations against the dead and non-infectious ones. Instead, metagenomics can possibly
176 narrow down the suite of contaminants to a selected few (e.g. Picornaviridae) for more in-depth
177 evaluation by other quantitative methods like quantitative PCR or digital PCR. By coupling
178 metagenomics with other quantitative methods, potential risks arising from recharging aquifer
179 with treated wastewater can be better assessed.

180
181

182 **Acknowledgements**

183 The authors would like to thank Dr Moustapha Harb and Dhafer Alsaleh for assistance during
184 the early sampling years. The authors would also like to thank Professor Tom Missimer and Mr.
185 Mohammed AlRayaan for providing background information related to the hydrogeological
186 characteristics of Wadi Fatimah. This study is funded by KAUST Baseline Grant BAS/1/1033-
187 01-01 awarded to P.-Y. Hong.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

Table 1. Summary of the organic matter components in Wadi Fatimah and upstream controls. DOC denotes dissolved organic carbon, BP denotes biopolymers, HS denotes humic substances, BB denotes building blocks, LMW denotes low molecular weight. Statistical comparison was made using t-test between Wadi Fatimah samples and upstream controls. ** denotes significant difference at 95% confidence level. * denotes significant difference at 90% confidence level. Value after \pm denotes standard deviation.

		Wadi Fatimah			Upstream controls	P-value
		March	August	December		
		2016	2016	2016		
DOC	mg/L	8.59 \pm 3.23	4.70 \pm 0.89	6.12 \pm 4.43	1.71 \pm 2.1	0.08*
Carbon distribution						
BP	mg/L	0.08 \pm 0.001	0.04 \pm 0.01	0.07 \pm 0.01	0.02 \pm 0.05	0.17
HS	mg/L	0.70 \pm 0.07	0.62 \pm 0.02	0.54 \pm 0.04	0.04 \pm 0.05	0.01**
BB	mg/L	3.20 \pm 0.15	2.58 \pm 0.34	2.11 \pm 0.09	0.68 \pm 0.88	0.08*
LMW neutrals	mg/L	3.36 \pm 0.33	0.58 \pm 0.06	1.28 \pm 0.11	0.14 \pm 0.06	0.33
LMW acids	mg/L	1.13 \pm 0.107	0.64 \pm 0.04	1.56 \pm 0.23	1.16 \pm 1.64	0.95

197 **Table 2.** Antibiotic resistance genes detected in groundwater samples. * denotes that the average percentage matched reads were
 198 already corrected for the average percentage detected in MilliQ control. N.D. denotes not detected or below detection limits.

Top 5 identified ARGs unique to Wadi Fatimah	Average % total matches *	Average % in MilliQ control	Denoted function
Aminoglycoside <i>aph3' Ia</i>	0.03	N.D.	Aminoglycoside phosphotransferase gene, confers resistance by catalyzing the addition of phosphate from ATP to the 3'-hydroxyl group of a 4,6-disubstituted aminoglycoside
Aminoglycoside <i>aadA13</i>	0.04	N.D.	Aminoglycoside nucleotidyltransferase gene, confers resistance by adenylyating the 2"-hydroxyl group of aminoglycoside
Beta-lactam-resistance <i>bla_{CAU-1}</i>	0.04	N.D.	Metallo-β-lactamase gene, confers resistance by catalyzing the hydrolysis of beta-lactam
Macrolide <i>ereA</i>	0.05	0.01	Erythromycin esterase gene, confers resistance through inactivation by hydrolyzing the lactone ring of the antibiotic
Plasmid-mediated-quinolone-resistance <i>qnrD1</i>	0.04		Quinolone resistance gene, confers resistance by preventing antibiotic binding
Top 5 identified ARGs unique to upstream controls	Average % total matches *	Average % in MilliQ control	Denoted function
Multidrug efflux pump <i>mdtE</i>	0.14	0.15	Part of the tripartite efflux system MdtEF-TolC, which confers resistance to compounds such as erythromycin, doxorubicin, benzalkonium and so on
Repressor for multidrug efflux pump <i>emrR</i>	0.12	0.13	Negatively regulates the transcription of the multidrug resistance pump-encoding operon, <i>emrRAB</i> , by binding to its regulatory region
Sensor kinase <i>cpxA</i>	0.11	0.13	Histidine kinase member of the two-component regulatory system CpxA/CpxR, responds to envelope stress response
Beta-lactam-resistance <i>bla_{OXA 281}</i>	0.10	N.D.	Oxacillinase (enzyme that confers resistance to oxacillin, a type of beta-lactam), presumably intrinsic to <i>Acinetobacter johnsonii</i> SH046
Beta-lactam-resistance <i>bla_{OXA 211}</i>	0.09	N.D.	Oxacillinase (enzyme that confers resistance to oxacillin, a type of beta-lactam), presumably intrinsic to <i>Acinetobacter johnsonii</i> SH046

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

ARGs identified in Wadi Fatimah and upstream controls	Average % total matches in Wadi Fatimah*	Average % total matches in upstream control*	Average % in MilliQ control	Denoted function
Beta-lactam-resistance <i>bla</i> _{OXA-212}	0.04	0.10	N.D.	Carbapenem-hydrolyzing class D beta-lactamase, confers resistance by catalyzing the hydrolysis of beta-lactam
Beta-lactam-resistance <i>bla</i> _{OXA-373}	0.03	0.14	N.D.	Functions as a transpeptidase, beta-lactam became suicide substrates of transpeptidases
Tetracycline <i>tet39</i>	0.04	0.16	0.05	Tetracycline resistance gene, confers resistance by active efflux mechanisms
Tetracycline <i>tetRG</i>	0.05	0.09	0.05	TetR is the repressor of the tetracycline resistance element, TetG is tetracycline efflux protein
Trimethoprim <i>dfr32</i>	0.05	0.03	0.12	Dihydrofolate reductase and trimethoprim resistance gene, confers resistance by replacing or substituting antibiotic action target

199

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

200 **Figure legends**

201 **Figure 1.** Bacterial communities and bacteria-associated antibiotic resistome in groundwater. **(a)**
202 Heat map illustrates relative abundance of individual bacteria phyla in the negative controls
203 (MilliQ water), upstream controls and wells in Wadi Fatimah across different sampling months.
204 **(b)** Bacteria alpha diversity in groundwater sampled from Wadi Fatimah across different
205 sampling months, upstream controls and negative controls. Linear line depicts an increasing
206 trend in alpha diversity with time. **(c)** Venn diagram illustrating the number of unique antibiotic
207 resistance genes detected in Wadi Fatimah (depicted as green) and upstream control samples
208 (depicted as red). Overlapping region shows the number of shared antibiotic resistance genes
209 detected in both Wadi Fatimah and upstream controls.

210 **Figure 2.** Viral communities in groundwater. **(a)** Heat map illustrates the relative abundance of
211 individual viral families in the control (MilliQ water), upstream controls and wells in Wadi
212 Fatimah across the different sampling times. **(b)** Virus alpha diversity in groundwater sampled
213 from Wadi Fatimah across different sampling months, upstream controls and negative controls.
214 Linear line depicts no apparent trend of increment with time.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

218 **References**

219 Al-Jassim, N., et al., 2015. Removal of bacterial contaminants and antibiotic resistance genes by
220 conventional wastewater treatment processes in Saudi Arabia: Is the treated wastewater safe to
221 reuse for agricultural irrigation? *Water Res.* 73, 277-90.

222 Connelly, S., et al., 2018. Distinct consequences of amoxicillin and ertapenem exposure in the porcine
223 gut microbiome. *Anaerobe.* 53, 82-93.

224 Grainger, D. J., 1992. A geological excursion in the Wadi Fatima area, near Jiddah, Saudi Arabia. *Geology*
225 *Today.* 8, 215-219.

226 Halliday, E., et al., 2014. Comparison of bacterial communities in sands and water at beaches with
227 bacterial water quality violations. *PLoS One.* 9, e90815.

228 Hasan, N. A., et al., 2014. Microbial Community Profiling of Human Saliva Using Shotgun Metagenomic
229 Sequencing. *Plos One.* 9.

230 Jumat, M. R., et al., 2017. Membrane Bioreactor-Based Wastewater Treatment Plant in Saudi Arabia:
231 Reduction of Viral Diversity, Load, and Infectious Capacity. *Water.* 9.

232 Kaleko, M., et al., 2016. Development of SYN-004, an oral beta-lactamase treatment to protect the gut
233 microbiome from antibiotic-mediated damage and prevent *Clostridium difficile* infection.
234 *Anaerobe.* 41, 58-67.

235 Luo, T., et al., Aqueduct Projected Water Stress Country Rankings. 2015.
236 [http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%](http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%20STRESS%20Country%20Rankings.pdf)
237 [20STRESS%20Country%20Rankings.pdf](http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%20STRESS%20Country%20Rankings.pdf). 31 August 2019

238 Manti, A., et al., 2008. Bacterial cell monitoring in wastewater treatment plants by flow cytometry.
239 *Water Environ Res.* 80, 346-54.

240 Ponnusamy, D., et al., 2016. Cross-talk among flesh-eating *Aeromonas hydrophila* strains in mixed
241 infection leading to necrotizing fasciitis. *Proceedings of the National Academy of Sciences.* 113,
242 722-727.

243 Quince, C., et al., 2017. Shotgun metagenomics, from sampling to analysis. *Nature Biotechnology.* 35,
244 833.

245 Reuter, G., et al., 2017. Saliviruses-the first knowledge about a newly discovered human picornavirus.
246 *Rev Med Virol.* 27.

247 Van Goethem, M. W., et al., 2018. A reservoir of 'historical' antibiotic resistance genes in remote pristine
248 Antarctic soils. *Microbiome.* 6, 40.

249 Versporten, A., et al., 2018. Antimicrobial consumption and resistance in adult hospital inpatients in 53
250 countries: results of an internet-based global point prevalence survey. *Lancet Glob Health.* 6,
251 e619-e629.

252 Zhang, K., Farahbakhsh, K., 2007. Removal of native coliphages and coliform bacteria from municipal
253 wastewater by various wastewater treatment processes: implications to water reuse. *Water*
254 *Res.* 41, 2816-24.

1
2
3
4 **Metagenomics-based evaluation of groundwater microbial profiles in response to treated**
5
6 **wastewater discharge**
7
8
9

10
11
12
13
14
15 Noor Zaouri ¹, Muhammad Raihan Jumat ², Tariq Cheema ³ and Pei-Ying Hong ^{1*}
16
17
18
19
20
21
22
23
24

25 ¹ Water Desalination and Reuse Center (WDRC), Biological and Environmental Sciences and
26 Engineering Division (BESE), King Abdullah University of Science and Technology (KAUST),
27 Thuwal-Jeddah 23955-6900, Saudi Arabia
28
29
30
31

32 ² Duke-NUS Graduate Medical School, 8 College Road, Singapore 169857
33
34

35 ³ Hydrology Department, Faculty of Earth Sciences, King Abdulaziz University (KAU), Jeddah,
36 Saudi Arabia
37
38
39
40
41
42
43

44 * Corresponding author:
45
46

47 Pei-Ying Hong
48

49 Tel: +966-(12)-808-2218;
50

51 Email: peiyong.hong@kaust.edu.sa
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Highlights

- Metagenomics as a tool to assess groundwater quality
- Lower alpha diversity of bacteria and virus due to treated wastewater exposure
- Increase in relative abundance of Planctomycetes and Picornaviridae
- Antibiotic resistance genes may not be good markers to differentiate wastewater exposure

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

2 **Abstract**

3 This study aims to demonstrate the use of metagenomics to assess groundwater quality.
4 Metagenomics revealed a lower alpha diversity for both bacteria and virus in wastewater-
5 exposed groundwater compared to the upstream controls. However, an increase in the relative
6 abundance of Planctomycetes and Picornaviridae, as well as increasing counts of unique
7 antibiotic resistance genes, was observed in wastewater-exposed groundwater. Findings suggest
8 that metagenomics can detect microbial signatures indicative of treated wastewater discharge.

9
10
11
12

13 **Keywords:** Shotgun metagenomics; aquifer recharge; groundwater quality; water reuse

1. Introduction

Water scarcity is projected to exacerbate by 2040, with 33 countries including Saudi Arabia predicted to face extreme water scarcity by that time (Luo et al., 2015). In recent years, recharging aquifer with treated wastewater has been proposed as a mitigation strategy to replenish the groundwater supplies. In Saudi Arabia, most wastewater treatment plants (WWTPs) are designed to achieve secondary or tertiary treatment with a final chlorination step prior to discharge. Such WWTPs typically achieve a cumulative 6-log reduction of microorganisms (Al-Jassim et al., 2015). Considering that untreated wastewater generally has a cell density of 10^{11} to 10^{12} cells/L (Manti et al., 2008), the final treated wastewater still contains a significant microbial load that may potentially perturb the indigenous groundwater microbial community. Aquifers are sensitive ecosystems, and therefore recharging aquifer with treated wastewater can only be adopted with an improved understanding on how these treated wastewater would affect the existing groundwater.

The current approaches to detecting changes in microbial community include qualitative approaches like terminal restriction fragment length polymorphism (T-RFLP) or quantitative approaches like real time PCR (qPCR). T-RFLP generates fingerprinting profiles for comparative analysis but does not provide information on the taxonomic identities of microbial contaminants. qPCR is a targeted approach that does not provide a comprehensive outlook on the wide variety of bacteria, virus and genes contaminants present in treated wastewater. The advent of high throughput untargeted sequencing approaches (e.g. metagenomics) have allowed a simultaneous evaluation on the bacteria, virus and genes present in an environmental sample (Quince et al., 2017). By comparing the diversity and relative abundance of these biological contaminants, one can denote if the groundwater has been perturbed by the treated wastewater.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

38 This study aims to demonstrate the use of metagenomics as a tool to assess microbial
39 water quality. Changes in the diversity and relative abundance of bacteria, virus and antibiotic
40 resistance genes were monitored to denote the extent of perturbation imposed by treated
41 wastewater.

42
43 **2. Material and method**

44 **2.1. Description of sampling sites and sampling procedure**

45 A subsurface reservoir, otherwise also known as Wadi Fatimah in Arabic, situated along the
46 Yamaniyah-Az Zibarah-Fatimah drainage system in Saudi Arabia (Grainger, 1992) was
47 identified for groundwater monitoring. Wadi Fatimah has been receiving treated wastewaters as
48 part of a study commissioned in 2010-2016 by Makkah water utilities to recharge local aquifers.
49 Groundwater at this site was collected over a period of 9 months, and include the following:
50 wells 1, 15, 14 and 17 in March 2016; wells 1, 14 and 17 in August 2016; wells 12, 15 and 17 in
51 December 2016 (Figure S1). Treated wastewaters injected to Wadi Fatimah were from a local
52 WWTP upstream of well 1, and utilized schematics including clarifiers, activated sludge tanks,
53 oxidation ponds and chlorine disinfection. Sampling wells varied throughout the sampling period
54 because of various reasons including lack of sufficient water to facilitate representative
55 sampling, groundwater well was closed for access or no permission was granted from the well
56 owners on the day of sampling. Two wells situated approximately 60 km upstream of Wadi
57 Fatimah were also identified, and groundwater was collected from these wells during the same
58 time frame as mentioned above. Samples from these two wells serve as upstream controls since
59 both the wells are not known to actively receive any treated wastewater discharge (Figure S1).
60 The collected volume of all groundwater samples ranged from 20-40 L per well, and were

1
2
3
4 61 comprised of a single grab sample collected per well per sampling event. Before collection, the
5
6 62 water was flushed for > 5 min to remove stagnant pools of water retained within the sampling
7
8
9 63 pipes. Water was also flushed until no apparent turbidity was observed prior to collection in
10
11
12 64 sterile polyethylene bottles. All samples were placed into a cooler and kept cool during transport.
13
14 65 Upon arrival in laboratory, samples were stored in 4 °C prior to analyses.

17 66 **2.2. Groundwater concentration and DNA extraction**

20 67 Each 20-40 L of groundwater sample was passed through a filtration system fitted with a T-
21
22
23 68 series cassette with a 100-kDa cut-off (approximately 0.005 µm) (Pall Corporation, Port
24
25 69 Washington, NY, USA). This cassette has been shown previously to retain viruses at an
26
27
28 70 efficiency of up to 87% and was particularly suited for large volume concentration (Jumat et al.,
29
30 71 2017). Hence, this procedure was chosen to concentrate any microbial matter present in the low
31
32
33 72 turbidity groundwater samples such as that collected in this study. Each groundwater sample was
34
35 73 concentrated to 5 mL of retentate. To collect any microbial matter adhering to the 100 kDa
36
37 74 cassette, the system was washed with 50 mL of wash solution (1 x Phosphate Buffered Saline
38
39
40 75 (PBS) and 0.01% Tween 60). After circulating the wash solution through the filtration system for
41
42 76 several minutes, the wash solution was itself concentrated to 5 mL. Both retentate and wash
43
44
45 77 solutions derived from the same sample were extracted for DNA was by using the DNeasy Blood
46
47 78 and Tissue Kit (Qiagen, Hilgen, Germany). MilliQ water was collected and subjected to the same
48
49
50 79 filtration procedure at the same time when all groundwater samples were filtered. This acts as a
51
52 80 negative control, and provide comparative analysis that would allow determination of microbial
53
54
55 81 populations unique to the groundwater.

56
57
58 82

83

84 **2.3. Metagenomic sequencing and analysis**

85 Genomic DNA was sent to CosmosID (Rockville, MD, USA) for sequencing on HiSeq, as
86 described (Hasan et al., 2014). Shotgun metagenomic datasets were directly analyzed by
87 CosmosID metagenomic software (CosmosID Inc., Rockville, MD) as described elsewhere
88 (Connelly et al., 2018; Hasan et al., 2014; Kaleko et al., 2016; Ponnusamy et al., 2016) to reveal
89 associated microbial community composition. Briefly, the system utilizes a high performance
90 data-mining k-mer algorithm and highly curated dynamic comparator databases (GenBook®)
91 that rapidly disambiguate millions of short reads into the discrete genomes or genes engendering
92 the particular sequences. The GeneBook databases were composed of over 150,000 microbial
93 genomes and gene sequences representing over 1000 bacterial and 5000 viral species, as well as
94 over 5500 antibiotic resistance genes. For each sample, the reads from a bacteria or viral species
95 are assigned to the strain with the highest aggregation statistics, and subsequently determined for
96 the corresponding alpha diversity. Community resistome was also identified using the CosmosID
97 Inc. bioinformatics software package by querying the unassembled sequence reads against the
98 CosmosID curated antibiotic resistance gene databases. Paired end Fastq files are deposited onto
99 European Nucleotide Archive under accession number PRJEB32173.

100 **2.4. Determination of organic water quality**

101 Organic constituents of the DOC were fractionated according to molecular weight (MW) with
102 size exclusion chromatography. The fraction distribution of the DOC in the groundwater was
103 analyzed using Liquid Chromatography equipped with a macro-porous silica separation column,
104 an Organic Carbon Detector (OCD), a UV 254 nm Detector (UVD) and an Organic Nitrogen

1
2
3
4 105 Detector (OND) (LC-OCD Model 8, DOC LABOR, Germany). ChromCaluni software was used
5
6
7 106 to determine the concentration of each fraction by integrating the area underneath the curve
8
9 107 according to manufacture instructions.

12 108 **3. Results and discussion**

15 109 **3.1. Taxonomic profile of bacterial communities**

16 110
17
18 111 Phyla Proteobacteria, Actinobacteria and Bacteroidetes constitute the three dominant bacterial
19
20 112 groups in the groundwater samples of both Wadi Fatimah and upstream controls. However,
21
22 113 Proteobacteria and Actinobacteria were also present in high relative abundance in the negative
23
24 114 MilliQ controls (Figure 1a). In contrast, Bacteroidetes was detected in higher relative abundance
25
26 115 and in more consistent frequency in only the groundwater samples. Planctomycetes was detected
27
28 116 in higher relative abundance among the Wadi Fatimah samples than that in upstream and
29
30 117 negative controls (Figure 1a). Planctomycetes has been shown to be closely associated with
31
32 118 organic detritus in the marine environment (Halliday et al., 2014). Similarly, a higher relative
33
34 119 abundance of Planctomycetes was observed in groundwater samples that had a higher dissolved
35
36 120 organic carbon (Table 1). Among these five detected DOC fractions, building block (BB) and
37
38 121 humic substances (HS) were significantly higher in groundwater sampled from Wadi Fatimah
39
40 122 compared to upstream controls ($p < 0.10$, Table 1). Specifically, HS which is predominated by
41
42 123 aromatic moieties was only detected in Wadi Fatimah at concentrations ranging from 0.54 to
43
44 124 0.70 mg/L. Despite the higher DOC concentrations in Wadi Fatimah, the alpha diversity of
45
46 125 bacterial community was lower in all of its samples compared to the controls (Figure 1b),
47
48 126 suggesting the presence of an unknown factor that bias against certain families (e.g.
49
50 127 Bradyrhizobiaceae, Methylobacteriaceae) in Wadi Fatimah. These families were only detected in
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2
3
4 128 the upstream controls during the sampling period. There was however, a positive trend of
5
6 129 increasing alpha diversity with time, although the reason behind this trend remains unknown.
7
8
9

10 130 **3.2. Community resistome**

11
12
13 131 Ninety six unique ARGs were detected in Wadi Fatimah while only 56 unique ARGs were
14
15
16 132 detected in the upstream controls (Figure 1c). The 5 most abundant unique ARGs detected in
17
18 133 Wadi Fatimah were related to aminoglycoside, beta-lactam, macrolide and quinolone resistance,
19
20
21 134 while the 5 most abundant ARGs detected in upstream controls also included resistance against
22
23 135 beta-lactam. Shared ARGs between Wadi Fatimah and control samples revealed additional genes
24
25
26 136 conferring resistance to tetracycline and trimethoprim (Table 2). A positive detection of ARGs
27
28 137 in both Wadi Fatimah and control samples suggests a ubiquitous presence of ARGs in most
29
30
31 138 environments, including pristine environment (Van Goethem et al., 2018). This may be related to
32
33 139 long term global consumption of antibiotics for both hospital-associated and community-
34
35 140 acquired infections (Versporten et al., 2018). Hence, the choice of ARGs as metagenomics
36
37
38 141 markers to denote anthropogenic contamination events may not be provide definitive results.
39
40

41 142 **3.3. Taxonomic profile of viral communities**

42
43
44 143 Podoviridae and Siphoviridae account as the two most dominant viral families in the
45
46 144 groundwater (Figure 2a). It was further observed that upstream controls had a significantly
47
48
49 145 higher relative abundance of Siphoviridae, along with Myoviridae, Inoviridae and Tectiviridae (p
50
51 146 < 0.10). All four families are mainly comprised of bacteriophages that rely on bacteria and
52
53
54 147 archaea as their natural hosts. Along with one of the negative controls which denoted an
55
56 148 exceedingly high relative abundance of Picornaviridae, Picornaviridae was consistently detected
57
58
59 149 in all groundwater samples and especially at a higher relative abundance in Wadi Fatimah
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

150 samples (average 45%) than in the upstream controls (average 11%). Picornaviridae comprises
151 of several notable human enteric viruses including Enteroviruses and Salivirus which are
152 proposed as potential human virus indicator (Reuter et al., 2017). The high relative abundance of
153 Picornaviridae may be related to the type of treatment process used by the WWTP situated in
154 Wadi Fatimah. Secondary and tertiary treatment processes typically do not achieve high removal
155 rates of viruses (Zhang and Farahbakhsh, 2007). In contrast, a WWTP retrofitted with membrane
156 bioreactor reported only sporadic occurrences of Enterovirus (Jumat et al., 2017). The alpha
157 diversity of viruses are higher in the controls than in Wadi Fatimah (Figure 2b). Coupled with
158 the higher abundance of families associated with bacteriophages, the higher alpha diversity may
159 be due to a wide variety of bacteriophages present in relatively unperturbed environments
160 compared to those exposed to treated wastewater. Unlike the bacterial alpha diversity, the viruses
161 did not exhibit an increasing trend of alpha diversity with time (Figure 2b).

162
163 **4. Conclusion**

164 Findings from this study demonstrates the use of metagenomics to provide a simultaneous
165 evaluation of bacteria, virus and antibiotic resistome in groundwater exposed to treated
166 wastewater discharge. It was observed that alpha diversity of both bacteria and viruses were
167 lower in exposed groundwater than the upstream controls. In addition, Planctomycetes and
168 Picornaviridae, both of which were detected by metagenomics simultaneously, may be
169 appropriate markers to indicate increased organic content and presence of human fecal
170 contamination. As more countries explore the use of treated wastewater to recharge diminishing
171 groundwater levels, metagenomics could suffice as an appropriate monitoring tool to ensure that
172 a wide diversity of biological contaminants are assessed and unintentional aberrations to the

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

173 groundwater minimized. However, metagenomics alone is unable to provide insights to the
174 potential risks that may arise from recharging aquifers with treated wastewater. This is because
175 metagenomics analysis does not provide absolute quantification of biological contaminants and
176 are derived from DNA template sequencing. Hence it is unable to differentiate the viable bacteria
177 populations against the dead and non-infectious ones. Instead, metagenomics can possibly
178 narrow down the suite of contaminants to a selected few (e.g. Picornaviridae) for more in-depth
179 evaluation by other quantitative methods like quantitative PCR or digital PCR. By coupling
180 metagenomics with other quantitative methods, potential risks arising from recharging aquifer
181 with treated wastewater can be better assessed.

182

183

Acknowledgements

185 The authors would like to thank Dr Moustapha Harb and Dhafer Alsaleh for assistance during
186 the early sampling years. The authors would also like to thank Professor Tom Missimer and Mr.
187 Mohammed AlRayaan for providing background information related to the hydrogeological
188 characteristics of Wadi Fatimah. This study is funded by KAUST Baseline Grant BAS/1/1033-
189 01-01 awarded to P.-Y. Hong.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

Table 1. Summary of the organic matter components in Wadi Fatimah and upstream controls. DOC denotes dissolved organic carbon, BP denotes biopolymers, HS denotes humic substances, BB denotes building blocks, LMW denotes low molecular weight. Statistical comparison was made using t-test between Wadi Fatimah samples and upstream controls. ** denotes significant difference at 95% confidence level. * denotes significant difference at 90% confidence level. Value after \pm denotes standard deviation.

		Wadi Fatimah			Upstream controls	P-value
		March	August	December		
		2016	2016	2016		
DOC	mg/L	8.59 \pm 3.23	4.70 \pm 0.89	6.12 \pm 4.43	1.71 \pm 2.1	0.08*
Carbon distribution						
BP	mg/L	0.08 \pm 0.001	0.04 \pm 0.01	0.07 \pm 0.01	0.02 \pm 0.05	0.17
HS	mg/L	0.70 \pm 0.07	0.62 \pm 0.02	0.54 \pm 0.04	0.04 \pm 0.05	0.01**
BB	mg/L	3.20 \pm 0.15	2.58 \pm 0.34	2.11 \pm 0.09	0.68 \pm 0.88	0.08*
LMW neutrals	mg/L	3.36 \pm 0.33	0.58 \pm 0.06	1.28 \pm 0.11	0.14 \pm 0.06	0.33
LMW acids	mg/L	1.13 \pm 0.107	0.64 \pm 0.04	1.56 \pm 0.23	1.16 \pm 1.64	0.95

199 **Table 2.** Antibiotic resistance genes detected in groundwater samples. * denotes that the average percentage matched reads were
 200 already corrected for the average percentage detected in MilliQ control. N.D. denotes not detected or below detection limits.

Top 5 identified ARGs unique to Wadi Fatimah	Average % total matches *	Average % in MilliQ control	Denoted function
Aminoglycoside <i>aph3' Ia</i>	0.03	N.D.	Aminoglycoside phosphotransferase gene, confers resistance by catalyzing the addition of phosphate from ATP to the 3'-hydroxyl group of a 4,6-disubstituted aminoglycoside
Aminoglycoside <i>aadA13</i>	0.04	N.D.	Aminoglycoside nucleotidyltransferase gene, confers resistance by adenylating the 2"-hydroxyl group of aminoglycoside
Beta-lactam-resistance <i>bla_{CAU-1}</i>	0.04	N.D.	Metallo-β-lactamase gene, confers resistance by catalyzing the hydrolysis of beta-lactam
Macrolide <i>ereA</i>	0.05	0.01	Erythromycin esterase gene, confers resistance through inactivation by hydrolyzing the lactone ring of the antibiotic
Plasmid-mediated-quinolone-resistance <i>qnrD1</i>	0.04		Quinolone resistance gene, confers resistance by preventing antibiotic binding
Top 5 identified ARGs unique to upstream controls	Average % total matches *	Average % in MilliQ control	Denoted function
Multidrug efflux pump <i>mdtE</i>	0.14	0.15	Part of the tripartite efflux system MdtEF-TolC, which confers resistance to compounds such as erythromycin, doxorubicin, benzalkonium and so on
Repressor for multidrug efflux pump <i>emrR</i>	0.12	0.13	Negatively regulates the transcription of the multidrug resistance pump-encoding operon, <i>emrRAB</i> , by binding to its regulatory region
Sensor kinase <i>cpxA</i>	0.11	0.13	Histidine kinase member of the two-component regulatory system CpxA/CpxR, responds to envelope stress response
Beta-lactam-resistance <i>bla_{OXA 281}</i>	0.10	N.D.	Oxacillinase (enzyme that confers resistance to oxacillin, a type of beta-lactam), presumably intrinsic to <i>Acinetobacter johnsonii</i> SH046
Beta-lactam-resistance <i>bla_{OXA 211}</i>	0.09	N.D.	Oxacillinase (enzyme that confers resistance to oxacillin, a type of beta-lactam), presumably intrinsic to <i>Acinetobacter johnsonii</i> SH046

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

ARGs identified in Wadi Fatimah and upstream controls	Average % total matches in Wadi Fatimah*	Average % total matches in upstream control*	Average % in MilliQ control	Denoted function
Beta-lactam-resistance <i>bla_{OXA-212}</i>	0.04	0.10	N.D.	Carbapenem-hydrolyzing class D beta-lactamase, confers resistance by catalyzing the hydrolysis of beta-lactam
Beta-lactam-resistance <i>bla_{OXA-373}</i>	0.03	0.14	N.D.	Functions as a transpeptidase, beta-lactam became suicide substrates of transpeptidases
Tetracycline <i>tet39</i>	0.04	0.16	0.05	Tetracycline resistance gene, confers resistance by active efflux mechanisms
Tetracycline <i>tetRG</i>	0.05	0.09	0.05	TetR is the repressor of the tetracycline resistance element, TetG is tetracycline efflux protein
Trimethoprim <i>dfr32</i>	0.05	0.03	0.12	Dihydrofolate reductase and trimethoprim resistance gene, confers resistance by replacing or substituting antibiotic action target

201

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

202 **Figure legends**

203 **Figure 1.** Bacterial communities and bacteria-associated antibiotic resistome in groundwater. **(a)**
204 Heat map illustrates relative abundance of individual bacteria phyla in the negative controls
205 (MilliQ water), upstream controls and wells in Wadi Fatimah across different sampling months.
206 **(b)** Bacteria alpha diversity in groundwater sampled from Wadi Fatimah across different
207 sampling months, upstream controls and negative controls. Linear line depicts an increasing
208 trend in alpha diversity with time. **(c)** Venn diagram illustrating the number of unique antibiotic
209 resistance genes detected in Wadi Fatimah (depicted as green) and upstream control samples
210 (depicted as red). Overlapping region shows the number of shared antibiotic resistance genes
211 detected in both Wadi Fatimah and upstream controls.

212 **Figure 2.** Viral communities in groundwater. **(a)** Heat map illustrates the relative abundance of
213 individual viral families in the control (MilliQ water), upstream controls and wells in Wadi
214 Fatimah across the different sampling times. **(b)** Virus alpha diversity in groundwater sampled
215 from Wadi Fatimah across different sampling months, upstream controls and negative controls.
216 Linear line depicts no apparent trend of increment with time.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

220 **References**

- 221 Al-Jassim, N., et al., 2015. Removal of bacterial contaminants and antibiotic resistance genes by
222 conventional wastewater treatment processes in Saudi Arabia: Is the treated wastewater safe to
223 reuse for agricultural irrigation? *Water Res.* 73, 277-90.
- 224 Connelly, S., et al., 2018. Distinct consequences of amoxicillin and ertapenem exposure in the porcine
225 gut microbiome. *Anaerobe.* 53, 82-93.
- 226 Grainger, D. J., 1992. A geological excursion in the Wadi Fatima area, near Jiddah, Saudi Arabia. *Geology*
227 *Today.* 8, 215-219.
- 228 Halliday, E., et al., 2014. Comparison of bacterial communities in sands and water at beaches with
229 bacterial water quality violations. *PLoS One.* 9, e90815.
- 230 Hasan, N. A., et al., 2014. Microbial Community Profiling of Human Saliva Using Shotgun Metagenomic
231 Sequencing. *Plos One.* 9.
- 232 Jumat, M. R., et al., 2017. Membrane Bioreactor-Based Wastewater Treatment Plant in Saudi Arabia:
233 Reduction of Viral Diversity, Load, and Infectious Capacity. *Water.* 9.
- 234 Kaleko, M., et al., 2016. Development of SYN-004, an oral beta-lactamase treatment to protect the gut
235 microbiome from antibiotic-mediated damage and prevent *Clostridium difficile* infection.
236 *Anaerobe.* 41, 58-67.
- 237 Luo, T., et al., Aqueduct Projected Water Stress Country Rankings. 2015.
238 [http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%](http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%20STRESS%20Country%20Rankings.pdf)
239 [20STRESS%20Country%20Rankings.pdf](http://reliefweb.int/sites/reliefweb.int/files/resources/AQUEDUCT%20PROJECTED%20WATER%20STRESS%20Country%20Rankings.pdf). 31 August 2019
- 240 Manti, A., et al., 2008. Bacterial cell monitoring in wastewater treatment plants by flow cytometry.
241 *Water Environ Res.* 80, 346-54.
- 242 Ponnusamy, D., et al., 2016. Cross-talk among flesh-eating *Aeromonas hydrophila* strains in mixed
243 infection leading to necrotizing fasciitis. *Proceedings of the National Academy of Sciences.* 113,
244 722-727.
- 245 Quince, C., et al., 2017. Shotgun metagenomics, from sampling to analysis. *Nature Biotechnology.* 35,
246 833.
- 247 Reuter, G., et al., 2017. Saliviruses-the first knowledge about a newly discovered human picornavirus.
248 *Rev Med Virol.* 27.
- 249 Van Goethem, M. W., et al., 2018. A reservoir of 'historical' antibiotic resistance genes in remote pristine
250 Antarctic soils. *Microbiome.* 6, 40.
- 251 Versporten, A., et al., 2018. Antimicrobial consumption and resistance in adult hospital inpatients in 53
252 countries: results of an internet-based global point prevalence survey. *Lancet Glob Health.* 6,
253 e619-e629.
- 254 Zhang, K., Farahbakhsh, K., 2007. Removal of native coliphages and coliform bacteria from municipal
255 wastewater by various wastewater treatment processes: implications to water reuse. *Water*
256 *Res.* 41, 2816-24.

Figure 1
[Click here to download high resolution image](#)

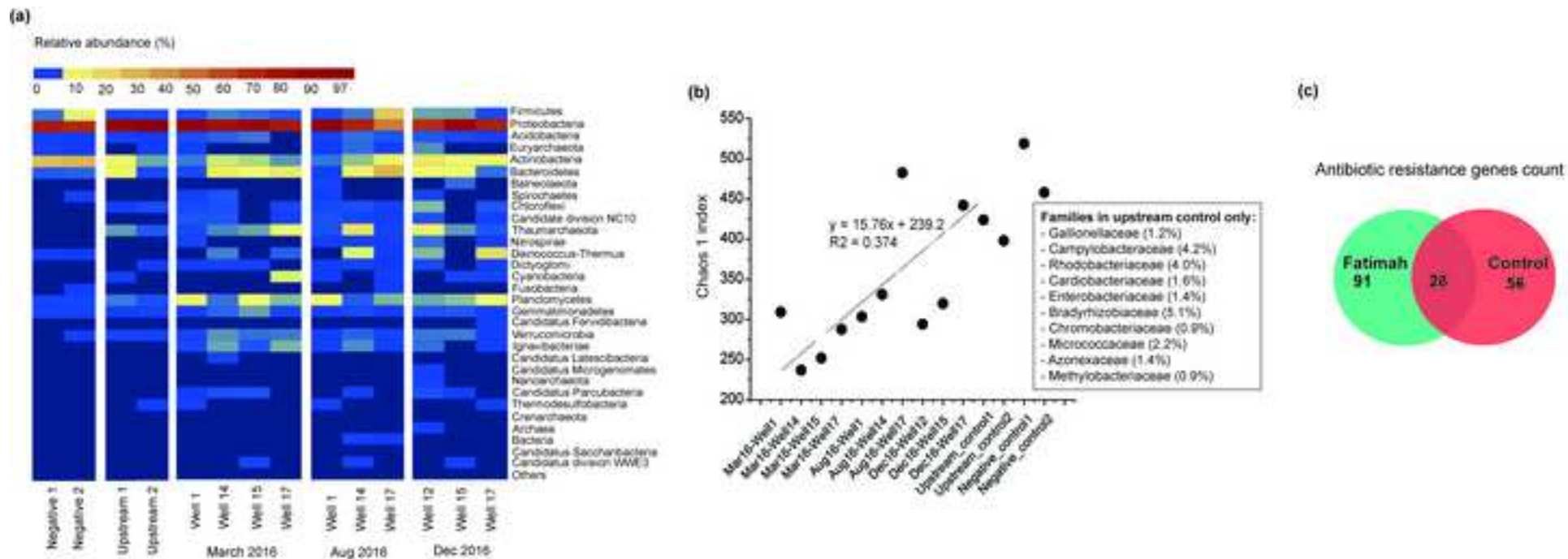
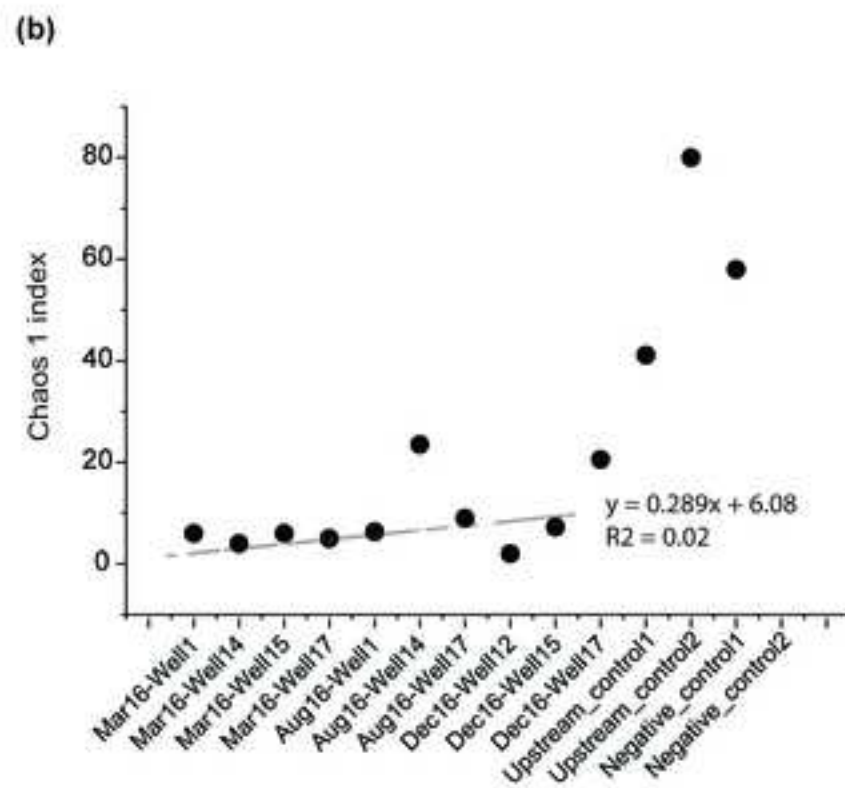
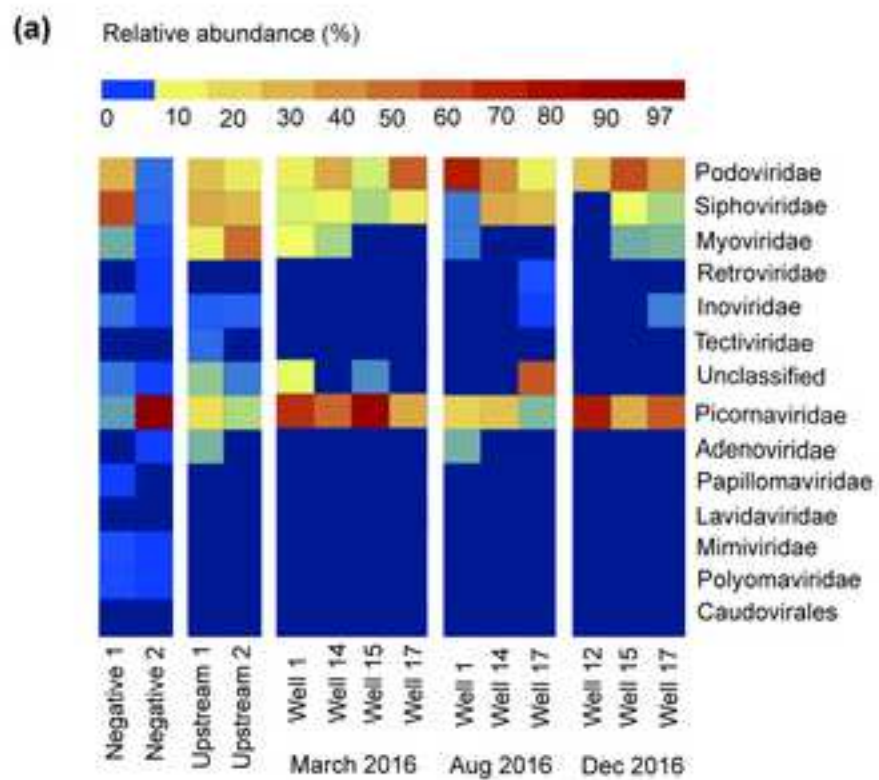


Figure 2
[Click here to download high resolution image](#)



Supplementary Material

[Click here to download Supplementary Material: For review - Supplemental Information.docx](#)

Conflict of interest

The authors declare no conflict of interest.