

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24

MS. SHANNON GRACE KLEIN (Orcid ID : 0000-0001-8190-3188)

Article type : Primary Research Articles

Title *Symbiodinium* mitigate the combined effects of hypoxia and acidification on a non-calcifying cnidarian

Running head CO₂ fuels a symbiotic cnidarian under hypoxia

Authors ¹Shannon G. Klein, ¹Kylie A. Pitt, ²Matthew R. Nitschke, ²Samantha Goyen, ³David T. Welsh, ²David J. Suggett, ³Anthony R. Carroll

Institute of Origin ¹Australian Rivers Institute – Coasts and Estuaries, Griffith School of Environment, Gold Coast, Griffith University, QLD 4222, Australia. ²Climate Change Cluster (C3), University of Technology Sydney, NSW 2007, Australia. ³Environmental Futures Research Institute, Griffith School of Environment, Gold Coast, Griffith University, QLD 4222, Australia.

Corresponding Author Shannon G. Klein, Ph: +61-7- 5552-8463, shannon.klein@griffithuni.edu.au

Keywords- Zooxanthellae, symbionts, photosynthesis, asexual reproduction, low DO, elevated pCO₂, low pH, jellyfish, *Cassiopea* sp.

Paper Type- Primary Research Article

Abstract

Anthropogenic nutrient inputs enhance microbial respiration within many coastal ecosystems, driving concurrent hypoxia and acidification. During photosynthesis, *Symbiodinium* spp., the

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1111/gcb.13718](https://doi.org/10.1111/gcb.13718)

This article is protected by copyright. All rights reserved

25 microalgal endosymbionts of cnidarians and other marine phyla, produce O₂ and assimilate CO₂,
26 and thus potentially mitigate the exposure of the host to these stresses. However, such a role for
27 *Symbiodinium* remains untested for non-calcifying cnidarians. We therefore contrasted the fitness
28 of symbiotic and aposymbiotic polyps of a model host jellyfish (*Cassiopea* sp.) under reduced O₂
29 (~2.09mgL⁻¹) and pH (~pH 7.63) scenarios in a full factorial experiment. Host fitness was
30 characterised as asexual reproduction and their ability to regulate internal pH and *Symbiodinium*
31 performance characterised by maximum photochemical efficiency, chl_a content, and cell density.
32 Acidification alone resulted in 58% more asexual reproduction of symbiotic polyps than
33 aposymbiotic polyps (and enhanced *Symbiodinium* cell density) suggesting *Cassiopea* sp. fitness
34 was enhanced by CO₂-stimulated *Symbiodinium* photosynthetic activity. Indeed, greater CO₂
35 drawdown (elevated pH) was observed within host tissues of symbiotic polyps under acidification
36 regardless of O₂ conditions. Hypoxia alone produced 22% fewer polyps than ambient conditions
37 regardless of acidification and symbiont status, suggesting *Symbiodinium* photosynthetic activity
38 did not mitigate its effects. Combined hypoxia and acidification, however, produced similar
39 numbers of symbiotic polyps compared with aposymbiotic kept under ambient conditions,
40 demonstrating that the presence of *Symbiodinium* was key for mitigating the combined effects of
41 hypoxia and acidification on asexual reproduction. We hypothesise that this mitigation occurred
42 because of reduced photorespiration under elevated CO₂ conditions where increased net O₂
43 production ameliorates oxygen debt. We show that *Symbiodinium* play an important role in
44 facilitating enhanced fitness of *Cassiopea* sp. polyps, and perhaps also other non-calcifying
45 cnidarian hosts, to the ubiquitous effects of ocean acidification. Importantly we highlight that
46 symbiotic, non-calcifying cnidarians may be particularly advantaged in productive coastal waters
47 that are subject to simultaneous hypoxia and acidification.

48 **Introduction**

49 Marine ecosystems are under increasing pressure from a suite of anthropogenic perturbations
50 (Crain *et al.*, 2008). Environmental hypoxia (reduced oxygen) is a particular threat to coastal
51 regions (Vaquer-Sunyer & Duarte, 2008, Chu & Tunnicliffe, 2015) and the ecosystem services
52 they sustain, such as fishery production (Breitburg *et al.*, 2009) and nutrient cycling (Woulds *et*
53 *al.*, 2007). Such hypoxic zones have been expanding since the mid-1900s via eutrophication
54 associated with heavily populated coastlines, and by 2008 hypoxia was estimated to have affected
55 more than 245,000 square kilometres of the Earth's surface (Diaz & Rosenberg, 2008). Hypoxic
56 zones are expected to further expand as ocean waters warm (Deutsch *et al.*, 2015, Schmidtko *et*
57 *al.*, 2017) and human populations become further concentrated along coastlines and catchments

58 (Rabalais *et al.*, 2010) and thus, hypoxia is considered one of the most severe threats to coastal
59 ecosystems.

60 Increased coastal hypoxia is primarily linked to accelerating rates of nutrient input along
61 coastlines (Vaquer-Sunyer & Duarte, 2008, Altieri & Gedan, 2015), which stimulate excessive
62 production of organic matter (OM) and microbial remineralisation processes to deplete oxygen
63 (O_2) from the water column (hypoxia, typically defined as $<2\text{mg } O_2 \text{ L}^{-1}$) (Cai *et al.*, 2011). Less
64 well recognised, however, is that hypoxic waters become, in parallel, more acidic since carbon
65 dioxide (CO_2) is simultaneously produced by respiration via microbial remineralisation, to reduce
66 pH through the formation and dissociation of carbonic acid (Gobler & Baumann, 2016). Such
67 hypoxia and associated acidification therefore become particularly amplified at night (Gobler &
68 Baumann, 2016) when photosynthesis ceases and community respiration increases (Baumann *et*
69 *al.*, 2015). Despite this inherent coupling of hypoxia and acidification, most experiments
70 investigating the potential impacts of hypoxia have not accurately replicated the water chemistry
71 associated with hypoxia. Instead, experiments commonly sparge seawater with N_2 gas (e.g. Wang
72 & Widdows, 1991, Baker & Mann, 1994, Gracey *et al.*, 2001, Eerkes-Medrano *et al.*, 2013),
73 which simultaneously displaces CO_2 and O_2 , to create conditions that are hypoxic but less (not
74 more) acidic. Furthermore, this approach is unable to replicate the inherent variation to
75 acidification and hypoxia that occurs on a diel basis (e.g. Regnault & Aldrich, 1988, Landry *et*
76 *al.*, 2007). Hence most manipulative experiments of hypoxia create conditions that are
77 inconsistent with typical natural conditions (Gobler *et al.*, 2014).

78 Understanding the biological responses to hypoxic conditions is fundamentally hindered
79 by limited data on the potential interactive effects of hypoxia and acidification on cnidarians (but
80 see, Steckbauer *et al.*, 2015). The few recent studies that have empirically examined the
81 interactive effects of hypoxia and acidification on marine invertebrates generally demonstrate
82 either additive (bivalves; Jakubowska & Normant, 2015, Jansson *et al.*, 2015) or synergistic
83 responses (gastropods, bivalves, anemones, respectively; Kim *et al.*, 2013, Gobler *et al.*, 2014,
84 Steckbauer *et al.*, 2015) to the dual stressors (but see, bivalves; Sui *et al.*, 2016). For example,
85 metabolism of two non-symbiotic, non-calcifying anemones (*Anemonia alicemartinae* and
86 *Phymactis papillosa*) was depressed under the combined effects of hypoxia and acidification but
87 increased under acidification in isolation (Steckbauer *et al.*, 2015). Taken together, these studies
88 generally suggest that the combined effects of hypoxia and acidification may be more severe than
89 those of the individual stressors.

90 Whilst many marine organisms appear to be negatively impacted by acidification (Kroeker
91 *et al.*, 2010, Przeslawski *et al.*, 2015), most observations of marine algae and seagrass suggest
92 productivity and competitive fitness will be either enhanced (or at worst generally unchanged,
93 Roleda *et al.*, 2012, Asnaghi *et al.*, 2013, Young & Gobler, 2016) by acidification. This poses an
94 interesting conundrum for how the diverse range of invertebrates that host symbiotic
95 dinoflagellates (e.g. *Symbiodinium* spp.) may respond to microbial-driven coastal hypoxia and
96 associated acidification. Specifically, the presence of *Symbiodinium* spp. may partially mitigate
97 the combined effects of hypoxia and acidification, where photosynthetic release of O₂ could
98 potentially ameliorate oxygen debt induced during hypoxic events (Malcolm & Brown, 1977)
99 whilst the simultaneous acquisition of CO₂ may reduce *p*CO₂ within their surrounding host cells
100 (Laurent *et al.*, 2013, Gibbin *et al.*, 2014, Laurent *et al.*, 2014). Indeed, photosynthetic activity of
101 *Symbiodinium* spp. may mitigate acidosis of host cells under high *p*CO₂ conditions (Gibbin *et al.*,
102 2014). Elevated *p*CO₂ of surrounding seawater may also stimulate the photosynthetic activity of
103 *Symbiodinium* spp. where cells have become CO₂ limited (see Suggett *et al.*, 2012, Suggett *et al.*,
104 2013, Gibbin *et al.*, 2014, Ventura *et al.*, 2016) and may act as a key condition needed to enhance
105 oxygenation of the host tissues. Whilst this potential mitigating role of endosymbiotic algae is
106 interesting, it may ultimately be restricted to certain *Symbiodinium* spp. genetic types with
107 inherently inefficient CO₂ acquisition modes under ‘present day’ seawater *p*CO₂ (see Brading *et*
108 *al.*, 2011, Brading *et al.*, 2013). Even so, the role of *in hospite* *Symbiodinium* spp. in potentially
109 mitigating the interactive effects of long-term hypoxia and acidification is unexplored.

110 Here we contrast the fitness of symbiotic and aposymbiotic polyps for a model host of
111 *Symbiodinium* spp. (the jellyfish *Cassiopea* sp., Hofmann *et al.*, 1978) in a full factorial design of
112 reduced O₂ and pH for the first time. Absolute values and extent of diel variability for pH and DO
113 were selected to mimic current-day hypoxic ecosystems. We specifically hypothesised that
114 *Cassiopea* sp. polyps exposed to hypoxia and acidification (either in isolation or combination)
115 would exhibit negative physiological effects but that these effects would be mitigated by the
116 presence of *Symbiodinium*. Here we define host fitness as the rate of asexual reproduction of
117 polyps and their ability to regulate internal pH (during the night and day), and the physiological
118 response of *Symbiodinium* spp. was characterised via measurements of maximum photochemical
119 efficiency, *chl a* content, and *Symbiodinium* cell density of symbiotic polyps. We also assessed
120 whether different *Symbiodinium* spp. genotypes were selected for under the different treatments.
121 From this experiment we present novel observations that demonstrate a potentially important role
122 of *Symbiodinium* in facilitating enhanced fitness of *Cassiopea* sp. polyps (and perhaps also other

123 non-calcifying cnidarians) to the ubiquitous effects of acidification and importantly that
124 *Symbiodinium* also appeared to sustain fitness of *Cassiopea* sp. polyps when acidification and
125 hypoxia co-occurred.

126 **Materials and methods**

127 *Species studied and response variables measured*

128 We examined both symbiotic and aposymbiotic (without symbionts) polyps of the upside down
129 jellyfish, *Cassiopea* sp. to test for the potential role of *Symbiodinium* in mitigating acidification/
130 hypoxia. *Cassiopea* sp. inhabit shallow tropical and sub-tropical coastal waters and lagoons
131 (Hofmann *et al.*, 1996) that exhibit considerable fluctuations of DO and pH (e.g. Gray *et al.*, 2012,
132 Tonetta *et al.*, 2014). *Cassiopea* sp. polyps were collected as larvae that had settled on a rock in a
133 display tank containing at least 10 adult medusae at *Underwater World*, Sunshine Coast,
134 Australia, in September 2013. Symbiotic polyps were sampled from the upper surface of the rock
135 and aposymbiotic polyps from the underside under low light conditions. *Cassiopea* spp. larvae are
136 aposymbiotic and metamorphose into aposymbiotic polyps (Sachs & Wilcox, 2006), which
137 subsequently acquire *Symbiodinium* cells from the external environment ('horizontal transmission',
138 Sachs & Wilcox, 2006, Thornhill *et al.*, 2006). Aposymbiotic and symbiotic *Cassiopea* sp. polyps
139 thus serve as ideal study organisms to examine for the role of *Symbiodinium* (and hence host-
140 *Symbiodinium* symbioses) in regulating host fitness.

141 *Experimental Approach*

142 Polyps were acclimated to laboratory conditions by maintaining them at 25°C (± 1 SE, 0.02)
143 under a 12:12 light: dark cycle of $\sim 470 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (Aqua Zonic: Super Actinic Blue
144 26W, 400-500nm and Super Sun 30W, 400-700nm) in fresh 10 μm filtered seawater that was
145 sourced from the Gold Coast Seaway, Queensland (27.56°S, 153.25°E). Seawater was stored in
146 darkness for at least 10 weeks prior to the start of the experiment to minimise exposure of polyps
147 to free-living *Symbiodinium* throughout experimentation. Polyps were fed newly hatched *Artemia*
148 sp. nauplii every third day. All polyps were checked for the presence or absence of symbionts
149 using pulse amplitude modulated (PAM) fluorometry (as used by: Steindler *et al.*, 2002, Lemloh
150 *et al.*, 2009) and fluorescence microscopy with UV illumination prior to the start of the
151 experiment and every second day during the experiment.

152 The experimental design consisted of three orthogonal factors: pH (control (24h mean
153 ± 1 SE, 7.95 ± 0.01) versus reduced (7.63 ± 0.01)), dissolved O₂ (DO) concentration (control (6.14

154 $\text{mgL}^{-1} \pm 0.03$) versus hypoxic ($2.09 \text{ mgL}^{-1} \pm 0.03$) and polyp type (symbiotic versus
155 aposymbiotic). Four replicate aquaria were randomly allocated to each combination of pH,
156 oxygen and polyp type (i.e. $n=32$). Six polyps were transferred using a toothpick into small plastic
157 petri dishes weighted with stainless steel weights. Three petri dishes were then immersed in 1L
158 glass aquaria (i.e. 18 polyps per aquarium); one petri dish was allocated for asexual reproduction
159 measurements and the other two were used for pH microelectrode and chlorophyll fluorescence
160 measurements. Three petri dishes were allocated to each replicate aquarium to prevent
161 overcrowding of polyps and to ensure that polyps used to measure one response variable were not
162 re-sampled for others during the experiment. The experiment ran for 22 days and was performed
163 in a controlled temperature laboratory with the ambient temperature set at 25°C .

164 Absolute values and extent of diel variability for pH and DO of control treatments were
165 replicated based on 24h field measurements taken in October, 2014 in Moreton Bay, Australia
166 (27.13°S , 153.07°E) (Fig. S1). In hypoxic systems, the magnitude of diel fluctuations of pH and
167 DO can vary depending on the ecosystem being tested. We, therefore, selected moderate levels of
168 DO and diel variation (1.5mgL^{-1} - 3.0mgL^{-1}) from a range data collected from current-day hypoxic
169 ecosystems in coastal ecosystems (e.g. Park *et al.*, 2007, Tyler *et al.*, 2009). Levels of DO and pH
170 are stoichiometrically linked in marine ecosystems (Cai *et al.*, 2011) and pH can vary between
171 6.9- 7.9 during hypoxic events (Gobler *et al.*, 2014, Gobler & Baumann, 2016). We, therefore,
172 selected moderate pH levels (pH 7.5-7.75) for the low pH treatments from a range of pH data
173 collected from coastal hypoxic systems (e.g. Cai *et al.*, 2011, Melzner *et al.*, 2012, Gobler *et al.*,
174 2014).

175 *Manipulation of water chemistry*

176 To achieve the desired water chemistry of each treatment, a series of gas mass controllers
177 were used to deliver mixtures of CO_2 , N_2 and O_2 gas to seawater (also see Bockmon *et al.*, 2013).
178 The desired gas compositions (CO_2 , N_2 , O_2) were mixed from individual gas cylinders using four
179 sets of three Omega® mass flow controllers (FMA-5400s, 0-20 mL/min (CO_2), 0-5 L/min (N_2),
180 0-2 L/min (O_2)), which allowed for four independent treatments. The mass flow controllers were
181 operated and functions monitored by a desktop PC running NI LabVIEW™ software (32-bit
182 version) with communication using a voltage generating Omega® Expandable Modular Data
183 Acquisition System® (iNET-400) connected with three Omega® wiring boxes with screw
184 terminals (iNET-510). The desired proportions of CO_2 , N_2 and O_2 were mixed in a stainless steel
185 manifold and the gas line that emerged from the manifold was split to provide identical gas

186 mixtures to the replicate aquaria. Gas flow rates to replicate aquaria were manually adjusted using
187 secondary stainless steel manifolds with control valves. For each treatment, two gas compositions
188 (day and night) were used to closely mimic diel fluctuations in water chemistry in the field. NI
189 LabVIEW™ was used to linearly transition between day and night gas mixtures but gas
190 compositions were held constant at night and from 10am-2pm (Fig. S2).

191 Desired gas compositions for each treatment were continuously delivered to each
192 aquarium using plastic air stones. Lids were placed loosely over each replicate aquarium with a
193 head space of ~10mm to minimise evaporation and subsequent changes in water chemistry. 25%
194 of the water for each aquarium was replaced every day using water of the same chemistry. All
195 aquaria were exposed to 12:12 light: dark cycle of ~470 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (Aqua Zonic: Super
196 Actinic Blue 26W, 400-500nm and Super Sun 30W, 400-700nm) throughout the experiment to
197 mimic diel patterns during summer.

198 *Analysis of carbonate chemistry*

199 Levels of $p\text{CO}_2$ were calculated based on measured levels of total alkalinity (TA), pH,
200 DO, temperature and salinity using the program CO₂SYS (Lewis *et al.*, 1998) (see Table S1).
201 Once per week, a 100mL water sample was collected for analysis of TA from one randomly
202 selected replicate from each of the treatments. Samples were collected in clean glass amber bottles
203 using a drawing tube. Bottles were filled from the bottom and water allowed to overflow for 10-15
204 seconds to minimise gas exchange with the atmosphere. All samples were fixed with 20 μL of
205 mercuric chloride to prevent biological activity and stored at 4 °C until analysed within 24 hours
206 of collection. TA samples were analysed using an automatic 848 Titrino Plus Total Alkalinity
207 Titrator (Metrohm®) calibrated every 3 days on the total scale using TRIS/HCl buffers in synthetic
208 seawater. TA measurements on 50mL samples of certified reference material (provided by A. G.
209 Dickson, batch #138) were used to verify TA values. Every third day, temperature, salinity, pH
210 and DO were measured at 10am (Table S1). Temperature was recorded in each aquarium using a
211 thermometer and salinity was measured using a conductivity-salinity metre (TPS salinity-
212 conductivity metre, MC-84). The DO concentration in each aquarium was recorded using an optic
213 DO sensor (Mettler Toledo OptiOPx, Mettler Toledo Ltd). The pH of each aquarium was
214 measured using a FiveGo pH meter (Mettler Toledo Ltd) equipped with a TRIS-compatible
215 electrode (Inlab Expert Pro Electrode, Mettler Toledo Ltd). Every 2-3 days, pH electrodes were
216 calibrated using TRIS/HCl buffers in synthetic seawater. To accurately measure diel variation in

217 O₂ and CO₂ during the experiment, pH and DO measurements were taken hourly (between 6am-
218 6pm) from one randomly selected replicate from each of the treatments once per week (Fig. S2).

219 *Asexual reproduction and internal pH measurements*

220 Asexual reproduction and internal pH (i.e. pH micro-electrode profiles) of both symbiotic
221 and aposymbiotic polyps were measured to investigate the extent to which *Symbiodinium* spp.
222 metabolism within host tissues mitigate external DO and/ or pH exposure. Every third day, polyp
223 dishes were removed from aquaria and checked for polyps undergoing strobilation (the production
224 of young medusae via transversal fission). At the end of the experiment, all reproduction dishes
225 were removed from aquaria and the number of individual polyps was recorded using a dissecting
226 microscope. Only asexual buds that had metamorphosed into individual polyps from the planuloid
227 stage were counted.

228 One polyp was selected randomly from each replicate aquarium to measure internal pH
229 every three days during the day and night to account for the lack of photosynthesis in symbiotic
230 polyps at night when acidification and hypoxia are amplified. All polyps were gently detached
231 from the polyp dishes allocated for pH profile measurements using stainless steel manicure
232 scissors and collected with ~30mL of their respective treatment water. If polyps were observed to
233 retract their oral arms, they were considered as 'stressed' and not used for microelectrode
234 measurements. Polyps were placed in a 30mm plastic petri dish under a dissecting microscope.
235 The pH microelectrode (pH-25 Unisense, Denmark, 20-30µm tip diameter) was mounted on, and
236 controlled by, a micromanipulator. *Cassiopea* sp. polyps are 'sticky' in texture and each polyp
237 was gently placed flush against the side of the petri dish to ensure the polyp would remain in a
238 fixed position. The pH microelectrode was introduced into the treatment water and external pH
239 measurements were taken ~5mm from the polyp. All microelectrode measurements were
240 performed horizontally through the polyp wall when polyps were alive and upright (i.e.
241 perpendicular to the microelectrode). The microelectrode was inserted halfway between the base
242 of the polyp and the beginning of the oral arms, to ensure consistency of microelectrode profiles
243 among all polyps. The epidermis formed a seal around the electrode and prevented fluid exchange
244 between the external seawater and the polyp tissue (*sensu* Revsbech *et al.*, 1995, Köhler-Rink &
245 Köhl, 2000). Polyps varied slightly in shape and size and so to ensure consistency, pH was
246 recorded continuously through one side of the polyp until the gastrovascular cavity was reached.
247 Four increments through the polyp were then determined relative to the thickness of the body wall
248 (i.e. the number of measurements taken through the polyp wall) to account for differences in

249 shape and size, yielding 6 measurements per polyp (i.e. external pH, increments 1-4 through the
250 polyp wall, and gastrovascular cavity). Prior to pH measurements, the pH micro-electrodes were
251 calibrated with pH 7 and 10 buffer solutions. Fluorescent pH sensitive dye, 5(6)-
252 Carboxynaphthofluorescein, (sourced from Sigma-Aldrich, CAS no: 128724-35-6) was used to
253 confirm relative pH changes of microelectrode profiles using a fluorescence microscope. Polyps
254 were noninvasively incubated at 25°C in their aquarium water supplemented with dye to a final
255 concentration of 50 $\mu\text{mol L}^{-1}$ 5(6)-Carboxynaphthofluorescein for 30 minutes to allow sufficient
256 uptake. Polyps were placed in 3mL vials and the pH-sensitive dye solution was allowed to
257 overflow the vials into water baths at a rate of $\sim 2\text{mL min}^{-1}$ to ensure the solution was gently
258 mixed. Following staining, polyps were gently placed on depression microscope slides using a
259 pipette and photographed horizontally using a Nikon Eclipse 80i fluorescence microscope with
260 UV illumination. The pH-sensitive dye solution was excited at a rate of 0.2Hz at $A_{598\text{nm}}$. To
261 confirm the response of the fluorescent dye, phosphate buffers of known concentrations were
262 supplemented with the pH-sensitive dye to a final concentration of 50 $\mu\text{mol L}^{-1}$ to construct a two-
263 point calibration curve. All polyps used for internal pH measurements were sacrificed.

264 *Maximum photochemical efficiency (F_v/F_m), chl a content and Symbiodinium cell density*

265 We measured another three response variables on symbiotic polyps (maximum
266 photochemical efficiency, F_v/F_m), Chl a content and *Symbiodinium* density) to assess potential
267 changes in *Symbiodinium* growth and photophysiological status in response to the various
268 treatments tested. Prior to the commencement of the experiment, and at weekly intervals
269 throughout the experiment, symbiotic polyps were sampled to measure maximum photochemical
270 efficiency of *Symbiodinium* using a Maxi-Imaging pulse amplitude modulator (Maxi-PAM, Walz
271 GmbH, Germany). Within each replicate aquarium, one polyp was transferred from the allocated
272 petri-dish for fluorescence measurements into an individual well of a black, non-binding 96-well
273 plate (Greiner Bio-One GmbH, cat no. 655090). All polyps were dark acclimated for 20 minutes
274 prior to PAM measurements. Repeated (n= 10, separated by 0.1s) chl a fluorescence inductions
275 were made to return values of the minimum (F_0) and maximum fluorescence yield (F_m), and
276 hence the maximum PSII photochemical efficiency ($F_v/F_m = [F_m - F_0]/F_m$).

277 At the end of the experiment, four polyps of similar shape and size were sampled from
278 each replicate aquarium containing symbiotic polyps and stored in 1mL of 0.22 μm filtered
279 seawater. Polyps were macerated with a tissue homogeniser for 30s and a 100 μL aliquot was
280 taken from each sample for *Symbiodinium* counts. 100 μL of glycerol was added to each 100 μL

281 aliquot and samples were frozen at -80°C until analysed. Defrosted samples were mixed and ten
282 $0.10\mu\text{L}$ drops from each sample were counted using a Neubauer haemocytometer. The remaining
283 $900\mu\text{L}$ samples were used for estimates of *chl a* content. All *chl a* samples were centrifuged at
284 $3000 \times g$ at 4°C for 10 minutes and the supernatant discarded. The pelleted *Symbiodinium* were
285 resuspended in 95% ethanol and extracted overnight in the dark at 4°C before centrifugation.
286 Absorption of the supernatant was then determined at 647nm and 664nm using a UV-1800
287 Shimadzu[®] spectrophotometer. All samples were analysed in 1cm quartz cuvettes and the
288 instrument was calibrated using 95% ethanol blanks. *Chla* content was determined using
289 coefficients from a spectrophotometric equation ($-2.6094 \times A_{629} + 12.4380 \times A_{665}$) for
290 dinoflagellates in ethanol (Ritchie, 2006). *Chla* (units) concentrations were normalised to
291 corresponding measures of cell content (units) to also yield the *chl a* $[\text{cell}]^{-1}$ (units). *Symbiodinium*
292 cell density and *chl a* concentrations were also normalised to the number of polyps analysed to
293 yield the *Symbiodinium* and *chl a* $[\text{polyp}]^{-1}$ (units), respectively.

294 *Identification of Symbiodinium genotype*

295 *Symbiodinium* genetic type identity was further determined across all (symbiotic)
296 treatments to confirm whether any changes associated with *Symbiodinium* reflected alterations in
297 physiology of the same type versus a switch in dominate type (e.g. Suggett *et al.*, 2012). At the
298 end of the experiment, ten polyps from each replicate aquarium containing *Symbiodinium* were
299 sampled and stored in DMSO preservation buffer (Seutin *et al.*, 1991). Samples were washed
300 twice with phosphate buffered saline (PBS) and the total DNA was extracted using the MO BIO
301 PowerPlant Pro DNA Isolation Kit (MO BIO Laboratories, CA, USA, cat no. 13400-50)
302 following the manufacturer's bead-beating protocol with an extra phenolic separation step. The
303 *Symbiodinium* partial 5.8S, ITS2, and partial 28S region was amplified by PCR using the forward
304 *ITS-dino* (5'GTGAATTGCAGAACTCCGTG 3') and reverse *ITS2-rev2*
305 (5'CCTCCGCTTACTTATATGCTT 3') primers (Stat *et al.*, 2011). ITS2 amplicons were purified
306 through gel electrophoresis, sequenced by the Australian Genome Research Facility (AGRF), and
307 compared to *Symbiodinium* entries in NCBI using the Basic Local Alignment Search Tool
308 (BLAST), yielding a % match. Sequences retrieved by this study were deposited in NCBI under
309 the accession numbers KX533944 through KX533954 (Table S3).

310 *Statistical analyses*

311 Dependent variables of number of polyps, *Chla*, *Symbiodinium* density and *Symbiodinium*
312 *Chla* per cell, were analysed using linear mixed models (LMMs) in SPSS (SPSS, Released 2013).

313 Prior to analyses, data were checked for normality and homoscedasticity using standardised
314 residual plots and Q-Q plots and, if required, data were either \ln or $\ln(x+1)$ transformed. All
315 factors were fixed and the number of factors for each dependent variable differed according to
316 how the data were collected. The factors for number of polyps were pH, oxygen concentration,
317 and polyp type, and for *Symbiodinium* Chla, cell density and chla [cell]⁻¹ were pH and oxygen
318 concentration. The dependent variables of internal pH and F_v/F_m were analysed using repeated
319 measures LMMs. The factors for internal pH were pH, oxygen concentration, polyp type, day/
320 night, and distance (through polyp), which was the repeated measure, whereas for F_v/F_m were pH,
321 oxygen concentration, and time, which was the repeated measure. In all repeated measures
322 LMMs, various models (e.g. AR(1), AR(1) heterogeneous, CS) were investigated to assess the
323 model of best fit by comparing several goodness-of-fit statistics (e.g. -2 Restricted Log
324 Likelihood, Akaike's Information Criteria (AIC) and Bayesian Information Criterion (BIC)).
325 Preliminary analysis for the dependent variable internal pH revealed no significant effect of the
326 factor oxygen for all terms, and thus this was removed and the analysis re-run. If significant
327 differences were found, estimated marginal means were used to determine which means differed.

328 **Results**

329 *Survival and asexual reproduction*

330 All polyps of *Cassiopea* sp. survived experimentation, with polyp numbers increasing via asexual
331 production in all treatments. At the end of the experiment, polyp numbers differed among pH
332 treatments but their response depended on symbiont status, resulting in a significant pH \times
333 symbiont interaction (Table 1, Fig. 1); specifically, greatest numbers were produced by symbiotic
334 polyps under low pH conditions, with symbiotic polyps producing 58% more than aposymbiotic
335 polyps. Symbiotic polyps still produced 17% more than aposymbiotic polyps under ambient pH.
336 Thus host fitness (asexual reproduction) was consistently higher when *Symbiodinium* was present
337 but the magnitude of the difference was greater in low pH conditions. Polyp numbers also differed
338 among oxygen treatments, with 22% fewer polyps occurring in hypoxic (mean \pm 1SE: $9.63 \pm$
339 0.53) compared to ambient treatments (12.25 ± 0.67), regardless of pH conditions or symbiont
340 status (Table 1). Overall, symbiotic polyps exposed to both low pH and low oxygen produced a
341 similar number of polyps (12.25 ± 0.41) as for symbiotic and aposymbiotic polyps exposed to
342 ambient conditions (13.25 ± 0.54 , 11.00 ± 0.70 , respectively) (Fig. S3). Thus the presence of
343 *Symbiodinium* appeared to enhance host fitness under acidification conditions that was otherwise
344 hindered under hypoxia alone. However, presence of *Symbiodinium* appeared to sustain host

345 fitness when hypoxia and acidification coincided. Strobilation was not observed during the
346 experiment.

347 *Internal pH profiles*

348 The pH profiles of polyp walls differed among pH treatments, symbiotic versus
349 aposymbiotic polyp, and between day and night, resulting in a significant day/night \times symbiont \times
350 pH \times distance interaction (Table 2, Fig. 2). During the night, pH profiles throughout the polyps
351 matched the external experimental conditions, whereby pH remained consistent across the
352 external to internal body wall and did not differ between symbiotic and aposymbiotic polyps (Fig.
353 2a). In contrast, during the day and for the symbiotic polyps, pH slowly increased from Increment
354 1 to Increment 4 through the polyp tissues and then decreased in the gastrovascular cavity to
355 levels similar to that of the surrounding water (Fig. 2b). Under ambient conditions, the internal pH
356 of symbiotic polyps was greater than aposymbiotic polyps and the internal pH of symbiotic polyps
357 increased by 0.19 units at Increment 4 relative to aposymbiotic polyps (Fig. 2b). As expected, this
358 pattern reflected the drawdown of inorganic carbon by *Symbiodinium* photosynthetic activity and
359 was also observed for the low pH treatments, but the magnitude of difference between symbiotic
360 and aposymbiotic polyps was greater under low pH conditions. The internal pH of symbiotic
361 polyps in the low pH treatment was highest at Increment 4 and increased by 0.27 units relative to
362 aposymbiotic polyps (Fig. 2b). At Increment 4, the internal pH of symbiotic polyps matched the
363 internal pH of aposymbiotic polyps under ambient conditions (Fig. 2b). There was no significant
364 difference between the pH of the gastrovascular cavities of symbiotic and aposymbiotic polyps,
365 but the pH levels of the gastrovascular cavity of polyps exposed to low pH conditions were
366 reduced (Fig. 2b).

367 *Maximum photochemical efficiency*

368 F_v/F_m remained generally consistent over time for all treatments (range: 0.143-0.325,
369 mean \pm 1SE: 0.258 \pm 0.003) and were similar to F_v/F_m values measured prior to the
370 commencement of the experiment (range: 0.265-0.338, mean \pm 1SE: 0.265 \pm 0.009). Although a
371 significant oxygen \times pH \times time interaction was detected (Table S2, Fig. S4), the magnitude of
372 difference between treatments was small and no consistent patterns were observed.

373 *Symbiodinium identification, Chla content and cell density*

374 *Symbiodinium* ITS2 sequences from all replicate aquaria, except for one replicate sample
375 in the control treatment that could not be sequenced, were confirmed to match *Symbiodinium* ITS2

376 subtype C1 (see NCBI Genbank accession KX533944 through KX533954 and Table S3 for ITS2
377 sequences generated in this study). Consequently, any change in fitness of symbiotic *Cassiopea*
378 sp. polyps reflects a change of the physiology (and/or cell density) of the existing *Symbiodinium*
379 type rather than a change towards alternate types with differing physiologies.

380 We subsequently examined *chl a* content and *Symbiodinium* cell density to evaluate how
381 the presence of symbionts potentially benefitted the host (as per Fig. 1). Specifically, at the end of
382 the experiment, *Symbiodinium* cell density polyp⁻¹ and the *Symbiodinium* specific *chl a* content
383 (units cell⁻¹) varied among pH treatments, but patterns were not consistent among oxygen
384 treatments (Table 3). Consistent with observations that symbiotic polyps produced more polyps
385 under low pH, *Symbiodinium* cell density in polyps in the low pH treatment was 39% higher than
386 those under ambient conditions and was higher than that of all other treatments (Fig. 3a).
387 However, this response was lost when low DO coincided with low pH, where polyps in the
388 hypoxic and low pH treatment had similar *Symbiodinium* cell densities to those in the control
389 treatment (Fig. 3a). Contrary to observations that low DO reduced host fitness (asexual
390 reproduction), polyps exposed to low DO had similar *Symbiodinium* cell densities to those in the
391 control treatment regardless of the pH conditions they were exposed to (Fig. 3a). Intriguingly,
392 observations of *Symbiodinium chl a* cell⁻¹ were not consistent with those of *Symbiodinium* cell
393 densities. *Symbiodinium chl a* cell⁻¹ was highest in the low pH and low DO treatment and exceeded
394 that of all other treatments (Fig. 3b). Consistent with reduced asexual reproduction of polyps
395 under low DO conditions, hypoxia in isolation resulted in a 52% lower *chl a* cell⁻¹ concentration
396 relative to the control treatment (Fig. 3b). This response was not consistent when symbiotic
397 polyps were exposed to hypoxia and low pH in combination, where *chl a* cell⁻¹ was 52% higher
398 under hypoxic and low pH conditions than under low pH conditions alone (Fig. 3b). **Discussion**

399 Elevated CO₂ (reduced pH) can benefit *Symbiodinium* both as free living cells (Brading *et al.*,
400 2011) and *in hospite* of cnidarians (e.g. anemones, Suggett *et al.*, 2012, Towanda & Thuesen,
401 2012; and corals, Crawley *et al.*, 2010, Suggett *et al.*, 2013). Our data are highly consistent with
402 these observations; specifically, that *Symbiodinium* spp. (ITS2 type C1) facilitated enhanced
403 fitness of *Cassiopea* sp. polyps under acidification conditions, whereby acidification alone
404 resulted in 58% more symbiotic polyps than aposymbiotic polyps and enhanced numbers of
405 *Symbiodinium* cells per polyp. Some studies, however, have reported no change (e.g. Brading *et*
406 *al.*, 2011) or even decreases (e.g. Anthony *et al.*, 2008) in rates of photosynthesis of
407 *Symbiodinium* under elevated CO₂ and thus, this effect may ultimately be restricted to certain
408 *Symbiodinium* spp. genetic types. Hypoxia alone reduced host fitness (asexual reproduction)

409 regardless of acidification and symbiont status, suggesting *Symbiodinium* photosynthetic activity
410 did not mitigate the negative effects of hypoxia. Most importantly, however, we observed that
411 hypoxia and acidification in combination produced as many symbiotic polyps as the aposymbiotic
412 polyps kept under ambient conditions. Hence, by enhancing photosynthetic activity, exposure to
413 elevated CO₂ appears to offset the negative effects of hypoxia in taxa that host *Symbiodinium*.
414 Our observations suggest that *Cassiopea* sp., and perhaps other symbiotic non-calcifying
415 cnidarians, may still thrive when hypoxia and acidification co-occur but non-symbiotic cnidarians
416 may be negatively impacted by the dual stressors. Species that host *Symbiodinium* only as
417 juveniles or adults, however, may still face challenges if their populations rely on recruitment of
418 aposymbiotic larvae.

419 Photosynthesis and photorespiration depend on the relative availability of O₂ and
420 dissolved inorganic carbon (DIC) (Larkum *et al.*, 2003, Crawley *et al.*, 2010) but no studies have
421 investigated the concurrent effects of hypoxia and acidification on these processes in symbionts.
422 Whilst our data cannot pinpoint the mechanism that appears to mitigate the observed negative
423 effects of hypoxia under acidification conditions, we hypothesise that our various observations
424 potentially indicate an important role for photorespiration upon exposure to the dual stressors. The
425 apparent mitigation of hypoxia only when acidification co-occurred in the presence of
426 *Symbiodinium in hospite* is consistent with reduced photorespiration under elevated CO₂
427 conditions (see Crawley *et al.*, 2010). A reduction in photorespiration would then increase net O₂
428 production and thus increase O₂ concentrations within host tissues thereby ameliorating oxygen
429 debt. Although data are limited on the effects of acidification on rates of photorespiration in
430 symbiotic cnidarians, some studies demonstrate increases in net oxygen production of non-
431 calcifying, symbiotic cnidarians (anemones, Suggett *et al.* 2012, Jarrold *et al.* 2013, Gibbin &
432 Davy 2014) under elevated pCO₂ conditions despite increased respiration rates at elevated pCO₂
433 concentrations (in some cases, Suggett *et al.* 2012, Gibbin & Davy 2014). Indeed, any increase in
434 net oxygen production must equate to a net increase in CO₂ fixation during photosynthesis (Reece
435 *et al.*, 2015). In addition to reduced costs of carbon acquisition under high pCO₂ (Ventura *et al.*,
436 2016), we hypothesise that inhibition of photorespiration may also partially explain the higher rate
437 of asexual reproduction of symbiotic polyps under acidification alone due to increased efficiency
438 of carbon fixation and increased availability of organic carbon for growth. Clearly, better
439 understanding processes such as photorespiration is needed to assess why only some
440 *Symbiodinium* types, i.e. perhaps those more susceptible to Rubisco oxygenation under relatively

441 low CO₂ conditions via differences in Rubisco pool sizes and/or turnover, appear to benefit from
442 elevated CO₂ availability (e.g. Brading *et al.*, 2013).

443 Survival of marine organisms in hypoxic systems may partly depend on their ability to
444 regulate their internal pH under the lowered pH conditions that result from elevated DIC
445 concentrations. During the day, under low pH conditions, the internal pH of symbiotic polyps
446 matched the pH levels of aposymbiotic polyps exposed to ambient conditions, which suggests that
447 photosynthesis of *Symbiodinium* regulated the internal pH of the polyps. Aposymbiotic polyps,
448 however, conformed to the pH of their respective treatments, indicating that *Cassiopea* sp. polyps,
449 as hosts, may have limited ability to regulate their internal pH. Our results are consistent with the
450 few other studies that have investigated the diel regulation of internal pH of cnidarian host cells
451 (Venn *et al.*, 2009, Laurent *et al.*, 2013, Laurent *et al.*, 2014, Gibbin *et al.*, 2014. For example, the
452 only other study that has compared the internal pH of symbiotic and non-symbiotic cells (isolated
453 from a coral) under elevated CO₂ conditions reported that the internal pH of non-symbiotic cells
454 decreased by 0.3-0.4 when exposed to decreasing pH (from 7.8 to 6.8) but the internal pH of
455 symbiotic cells recovered to control levels (Gibbin *et al.*, 2014). Combined, these studies highlight
456 the significant control that *Symbiodinium* exert over the internal pH of their host tissues and
457 suggest that symbiotic biota will typically be more robust to hypoxic environments than non-
458 symbiotic biota.

459 Survival of symbiotic organisms ultimately depends on the physiological limitations of
460 both the host and their symbionts. Maximum photochemical efficiency values of *Symbiodinium*
461 were low relative to those previously observed for *Symbiodinium in hospite* of cnidarians (e.g.
462 Enochs *et al.*, 2014, Hoadley *et al.*, 2015, including polyps of *Cassiopea* sp. Klein *et al.*, 2016)
463 and could reflect a number of biological (e.g. high light fields, chlororespiration), or measurement
464 artefacts (e.g. lower values expected with imaging PAM, Levin *et al.*, 2017) that cannot presently
465 be ascertained. Low maximum photochemical efficiency values were unlikely induced by stress
466 since all other response variables (i.e. asexual reproduction, *Symbiodinium* densities, Chla cell⁻¹)
467 indicate that ambient experimental conditions were optimal for *Symbiodinium*. Even so, maximum
468 photochemical efficiency appeared to be unaffected by the various treatments tested, suggesting
469 that any artefact was constant. *Symbiodinium* densities (and host asexual reproduction), however,
470 were highest in the low pH treatment, suggesting that low pH conditions are favourable for
471 *Symbiodinium*. Such a pattern is consistent with studies on anemones where elevated CO₂
472 (reduced pH) not only enhanced *Symbiodinium* (ITS2 type A19) numbers and productivity, but
473 natural population sizes were also substantially increased in proximity to a natural CO₂ vent

474 (Suggett *et al.*, 2012). As with this previous observation, ours similarly suggests that
475 *Symbiodinium* of *Cassiopea* sp. polyps may be limited by the availability of dissolved inorganic
476 carbon (DIC) under present day $p\text{CO}_2$ conditions. Exposure to low pH and low DO
477 simultaneously did not increase densities of *Symbiodinium*; however, $\text{chl}a$ cell^{-1} increased in
478 response to the dual stressors, consistent with observations that *Symbiodinium* resource
479 investment into pigment synthesis and/ or division rate appears carbon limited under ambient
480 conditions (e.g. ITS2 type A13, Brading *et al.*, 2011). No studies have examined the response of
481 *Symbiodinium* to the combined effects of hypoxia and acidification but observations of other non-
482 calcifying symbiotic cnidarians exposed to elevated CO_2 in isolation demonstrate inconsistent
483 responses of *Symbiodinium* cell densities and Chla content (e.g. Suggett *et al.*, 2012, Towanda &
484 Thuesen, 2012, Gibbin & Davy, 2014, Horwitz *et al.*, 2015). For example, *Symbiodinium* cell
485 densities and chlorophyll content of the anemone *A. elegantissima* were unaffected by high CO_2
486 conditions despite increased rates of photosynthesis (Horwitz *et al.*, 2015). However, consistent
487 with the current study, acidification increased symbiont densities in two anemones, *Anthopleura*
488 *elegantissima* and *Aiptasia* sp. (Suggett *et al.*, 2012, Gibbin & Davy, 2014). Together these results
489 further highlight the complexity of responses of cnidarian associations and demonstrate
490 differential effects of high CO_2 conditions among *Symbiodinium* genotypes.

491 Some cnidarians host multiple, genetically distinct variants of *Symbiodinium* that tolerate
492 different types or levels of environmental stress (Baker, 2003, Thornhill *et al.*, 2006, Putnam *et*
493 *al.*, 2012). Hosting multiple variants of *Symbiodinium* may confer a fitness benefit to the host,
494 particularly if the composition of the symbiont community varies (e.g. via competitive
495 displacement) in response to changing environmental conditions (so called ‘symbiont shuffling’)
496 (Little *et al.*, 2004, Berkelmans & van Oppen, 2006). However, in our study we only detected
497 *Symbiodinium* ITS2 type C1, which is a generalist symbiont (LaJeunesse, 2005) that has been
498 observed in >100 host species, including *Cassiopea* (Franklin *et al.*, 2012, Tonk *et al.*, 2013,
499 Mellas *et al.*, 2014). Since we detected only one variant of *Symbiodinium*, there was no evidence
500 of major symbiont shuffling of the dominant symbiont populations, although we cannot rule out
501 that the techniques used in this study and the ITS2 marker may not fully capture changes in
502 dominance of population heterogeneity, notably for type C1 (see Howells *et al.*, 2016, Wham &
503 LaJeunesse, 2016). It is also possible that cryptic *Symbiodinium* variants may have occurred at
504 levels below the detection thresholds of the techniques we used (e.g. Boulotte *et al.*, 2016),
505 although the short duration of our experiment (but see Lewis & Coffroth, 2004) probably
506 precluded the potential for shuffling to occur. Even if some *Cassiopea* sp. polyps harbour only

507 one *Symbiodinium* type, they may still change their symbionts via uptake of new symbiont types
508 from the environment (termed symbiont ‘switching’, *sensu* Baker, 2003). Whilst the typical
509 symbionts that are *in hospite* of *Cassiopea* spp. populations in Australia are at present unknown, it
510 is well demonstrated that *Cassiopea* spp. polyps can host multiple variants of *Symbiodinium* and
511 acquire additional symbiont types at the polyp stage (including clades A, B, C and D, Mellas *et*
512 *al.*, 2014). However, polyps in our experiment were not exposed to exogenous *Symbiodinium* cells
513 and thus could not have acquired new symbiont types unless symbionts were shared horizontally
514 between polyps (see Sachs & Wilcox, 2006). To more accurately predict how symbiotic
515 cnidarians, as a group, may respond to hypoxia and acidification we must also consider that hosts
516 may acquire more resistant symbionts, including types that are physiologically adapted to extreme
517 environmental conditions in the longer term (Brading *et al.*, 2011).

518 Various combinations of host and symbiont type may provide physiological advantages
519 under changing ocean conditions. Our observations suggest that *Cassiopea* sp. harbouring
520 *Symbiodinium* subclade C1 responded positively to acidification conditions. Although no studies
521 have investigated the combined effects of hypoxia and acidification on *in hospite Symbiodinium*,
522 our results are consistent with studies that investigated the future effects of ocean acidification on
523 other cnidarians that host *Symbiodinium* types. Indeed, high CO₂ conditions stimulated the
524 productivity of two anemones, *Anthopleura elegantissima* and *Anemonia viridis*, harbouring
525 *Symbiodinium* clade B and A19, respectively (Suggett *et al.*, 2012, Towanda & Thuesen, 2012).
526 To better determine how biota may respond to hypoxia and acidification, we must now determine
527 whether our results for *Symbiodinium* C1 and *Cassiopea* sp. polyps are consistent with other
528 symbiont types and hosts, and investigate possible interactions between other environmental
529 stressors. Indeed, *Cassiopea* spp. harbour other clades of *Symbiodinium* including A, B and D
530 (Santos *et al.*, 2002, Thornhill *et al.*, 2006, Mellas *et al.*, 2014), and whether our observations here
531 scale to other *Cassiopea* species, life history stages, and/or symbiont types remains to be tested.

532 Our understanding of the responses of marine biota to hypoxic conditions is hindered
533 because the majority of hypoxia studies manipulate O₂ levels with N₂ gas (Gobler & Baumann,
534 2016), thereby increasing (up to pH 8.6, see Gobler *et al.*, 2014) and not decreasing pH. In the
535 current study, the fitness of symbiotic polyps appeared to be enhanced by acidification under
536 hypoxic conditions, suggesting that studies that do not account for concurrent changes in O₂ and
537 CO₂ may produce results that do not accurately reflect the response of symbiotic biota to hypoxic
538 environments. Aposymbiotic polyps, however, were negatively affected by hypoxia regardless of
539 the pH conditions they were exposed to, suggesting that acidification did not exacerbate the

540 effects of low oxygen availability. Our observations of aposymbiotic polyps are inconsistent with
541 the only other study to examine the interactive effects of low DO and pH on non-symbiotic, non-
542 calcifying cnidarians (two species of anemones (*Anemonia alicemartinae* and *Phymactis*
543 *papillosa*) (Steckbauer *et al.*, 2015). Although both anemones are naturally non-symbiotic (unlike
544 *Cassiopea* sp.), the study demonstrated that exposure to acidification alone increased the
545 metabolism of *A. alicemartinae* and *P. papillosa* but exposure to acidification and hypoxia in
546 combination depressed metabolism of both species. However, we cannot determine whether our
547 results are consistent with other studies that mimic hypoxia using N₂ gas because we did not
548 expose aposymbiotic polyps to low DO and high pH in combination; for this, studies will need to
549 further compare the response of biota to low DO and high pH in combination to those exposed to
550 hypoxia in isolation to assess the reliability of results obtained by studies of hypoxia that
551 manipulate O₂ levels with N₂ gas.

552 *Symbiodinium* are clearly important in mitigating the combined effects of hypoxia and
553 acidification on *Cassiopea* sp. polyps. Our data suggest that symbiotic *Cassiopea* sp. may still
554 thrive in hypoxic environments and although aposymbiotic *Cassiopea* sp. may persist, they are
555 unlikely to proliferate when exposed to the dual stressors. Symbiotic (non-calcifying) cnidarians,
556 such as jellyfish (tested here) but perhaps also other non-calcifying cnidarians may therefore have
557 a greater competitive advantage in current-day hypoxic zones. Our observations that
558 *Symbiodinium* mitigated the negative effects of hypoxia when acidification co-occurred highlights
559 the importance of investigating the concurrent effects of hypoxia and acidification on symbiotic
560 biota. Although the response of aposymbiotic polyps to hypoxia was unaffected by acidification,
561 manipulative experiments of hypoxia need to consider concurrent changes in pH to accurately
562 reflect field observations of hypoxic zones. We therefore advocate for a prompt re-alignment of
563 future studies of hypoxia and suggest that future experiments consider concurrent changes in DO
564 and pH. Whilst *Symbiodinium* may benefit non-calcifying cnidarians in current-day hypoxic zones
565 (tested here), *Symbiodinium* are sensitive to transient heat stress and thus to understand whether
566 this potentially important role of *Symbiodinium* holds under future additional warming scenarios
567 we now need to investigate the effects of additional warming in combination with hypoxia and
568 acidification.

569 **Acknowledgements**

570 Funding for this study was provided by Griffith University and an Australian Post-graduate
571 Award to S.G.K. The contribution of D.J.S. and M.R.N. to this work was supported through an

572 Australian Research Council (ARC) Discovery Grant DP160100271. We thank W. Bennett, F.
573 Leusch and D. Tonzing for technical assistance and, J. Arthur and J. Hay for statistical advice. We
574 also thank A. Reno and K. Wilson from *Underwater World*, Sunshine Coast, Australia for cultures
575 of *Cassiopea* sp. polyps.

576 **References**

- 577 Altieri AH, Gedan KB (2015) Climate change and dead zones. *Global Change Biology*, **21**, 1395-
578 1406.
- 579 Anthony KR, Kline DI, Diaz-Pulido G, Dove S, Hoegh-Guldberg O (2008) Ocean acidification
580 causes bleaching and productivity loss in coral reef builders. *Proceedings of the National*
581 *Academy of Sciences*, **105**, 17442-17446.
- 582 Asnaghi V, Chiantore M, Mangialajo L, Gazeau F, Francour P, Alliouane S, Gattuso J-P (2013)
583 Cascading effects of ocean acidification in a rocky subtidal community. *PLOS one*, **8**,
584 e61978.
- 585 Baker AC (2003) Flexibility and specificity in coral-algal symbiosis: diversity, ecology, and
586 biogeography of *Symbiodinium*. *Annual Review of Ecology, Evolution, and Systematics*,
587 **34**, 661-689.
- 588 Baker S, Mann R (1994) Description of metamorphic phases in the oyster *Crassostrea virginica*
589 and effects of hypoxia on metamorphosis. *Marine Ecology Progress Series*, **104**, 91-91.
- 590 Baumann H, Wallace RB, Tagliaferri T, Gobler CJ (2015) Large natural pH, CO₂ and O₂
591 fluctuations in a temperate tidal salt marsh on diel, seasonal, and interannual time scales.
592 *Estuaries and Coasts*, **38**, 220-231.
- 593 Berkelmans R, Van Oppen MJ (2006) The role of zooxanthellae in the thermal tolerance of corals:
594 a 'nugget of hope' for coral reefs in an era of climate change. *Proceedings of the Royal*
595 *Society of London B: Biological Sciences*, **273**, 2305-2312.
- 596 Bockmon E, Frieder C, Navarro M, White-Kershek L, Dickson A (2013) Technical Note:
597 Controlled experimental aquarium system for multi-stressor investigation of carbonate
598 chemistry, oxygen saturation, and temperature. *Biogeosciences*, **10**, 5967-5975.
- 599 Boulotte NM, Dalton SJ, Carroll AG, Harrison PL, Putnam HM, Peplow LM, Van Oppen MJ
600 (2016) Exploring the *Symbiodinium* rare biosphere provides evidence for symbiont
601 switching in reef-building corals. *The ISME journal*, **10**, 2693-2701.
- 602 Brading P, Warner ME, Davey P, Smith DJ, Achterberg EP, Suggett DJ (2011) Differential
603 effects of ocean acidification on growth and photosynthesis among phylotypes of
604 *Symbiodinium* (Dinophyceae). *Limnology and Oceanography*, **56**, 927-938.

- 605 Brading P, Warner ME, Smith DJ, Suggett DJ (2013) Contrasting modes of inorganic carbon
606 acquisition amongst *Symbiodinium* (Dinophyceae) phylotypes. *New Phytologist*, **200**, 432-
607 442.
- 608 Breitburg DL, Hondorp DW, Davias LA, Diaz RJ (2009) Hypoxia, nitrogen, and fisheries:
609 integrating effects across local and global landscapes. *Marine Science*, **1**, 329-349.
- 610 Cai W-J, Hu X, Huang W-J *et al.* (2011) Acidification of subsurface coastal waters enhanced by
611 eutrophication. *Nature Geoscience*, **4**, 766-770.
- 612 Chu JW, Tunnicliffe V (2015) Oxygen limitations on marine animal distributions and the collapse
613 of epibenthic community structure during shoaling hypoxia. *Global Change Biology*, **21**,
614 2989-3004.
- 615 Crain CM, Kroeker K, Halpern BS (2008) Interactive and cumulative effects of multiple human
616 stressors in marine systems. *Ecology Letters*, **11**, 1304-1315.
- 617 Crawley A, Kline DI, Dunn S, Anthony KEN, Dove S (2010) The effect of ocean acidification on
618 symbiont photorespiration and productivity in *Acropora formosa*. *Global Change Biology*,
619 **16**, 851-863.
- 620 Deutsch C, Ferrel A, Seibel B, Pörtner H-O, Huey RB (2015) Climate change tightens a metabolic
621 constraint on marine habitats. *Science*, **348**, 1132-1135.
- 622 Diaz RJ, Rosenberg R (2008) Spreading dead zones and consequences for marine ecosystems.
623 *Science*, **321**, 926-929.
- 624 Eerkes-Medrano D, Menge BA, Sislak C, Langdon CJ (2013) Contrasting effects of hypoxic
625 conditions on survivorship of planktonic larvae of rocky intertidal invertebrates. *Marine*
626 *Ecology Progress Series*, **478**, 139-151.
- 627 Enochs IC, Manzello DP, Carlton R, Schopmeyer S, van Hooidonk R, Lirman D (2014) Effects of
628 light and elevated $p\text{CO}_2$ on the growth and photochemical efficiency of *Acropora*
629 *cervicornis*. *Coral Reefs*, **33**, 477-485.
- 630 Franklin EC, Stat M, Pochon X, Putnam HM, Gates RD (2012) GeoSymbio: a hybrid,
631 cloud-based web application of global geospatial bioinformatics and ecoinformatics for
632 *Symbiodinium*-host symbioses. *Molecular Ecology Resources*, **12**, 369-373.
- 633 Gibbin EM, Davy SK (2014) The photo-physiological response of a model cnidarian-
634 dinoflagellate symbiosis to CO_2 -induced acidification at the cellular level. *Journal of*
635 *Experimental Marine Biology and Ecology*, **457**, 1-7.
- 636 Gibbin EM, Putnam HM, Davy SK, Gates RD (2014) Intracellular pH and its response to CO_2 -
637 driven seawater acidification in symbiotic versus non-symbiotic coral cells. *Journal of*
638 *Experimental Biology*, **217**, 1963-1969.

- 639 Gobler CJ, Baumann H (2016) Hypoxia and acidification in ocean ecosystems: coupled dynamics
640 and effects on marine life. *Biology Letters*, **12**, 20150976.
- 641 Gobler CJ, Depasquale EL, Griffith AW, Baumann H (2014) Hypoxia and acidification have
642 additive and synergistic negative effects on the growth, survival, and metamorphosis of
643 early life stage bivalves. *PLOS one*, **9**, e83648.
- 644 Gracey AY, Troll JV, Somero GN (2001) Hypoxia-induced gene expression profiling in the
645 euryoxic fish *Gillichthys mirabilis*. *Proceedings of the National Academy of Sciences*, **98**,
646 1993-1998.
- 647 Gray SE, Degrandpre MD, Langdon C, Corredor JE (2012) Short-term and seasonal pH, $p\text{CO}_2$
648 and saturation state variability in a coral-reef ecosystem. *Global Biogeochemical Cycles*,
649 **26**, GB3012.
- 650 Hoadley KD, Pettay DT, Grottoli AG, Cai WJ, Melman TF, Schoepf V, Hu X, Li Q, Xu H, Wang
651 Y, Matsui Y, Baumann JH, Warner ME (2015) Physiological response to elevated
652 temperature and $p\text{CO}_2$ varies across four Pacific coral species: Understanding the unique
653 host+symbiont response. *Scientific Reports*, **5**, 18371.
- 654 Hofmann D, Neumann R, Henne K (1978) Strobilation, budding and initiation of scyphistoma
655 morphogenesis in the rhizostome *Cassiopea andromeda* (Cnidaria: Scyphozoa). *Marine
656 Biology*, **47**, 161-176.
- 657 Horwitz R, Borell EM, Yam R, Shemesh A, Fine M (2015) Natural high $p\text{CO}_2$ increases
658 autotrophy in *Anemonia viridis* (Anthozoa) as revealed from stable isotope (C, N) analysis.
659 *Scientific Reports*, **5**, 8779.
- 660 Howells E, Willis B, Bay L, Van Oppen M (2016) Microsatellite allele sizes alone are insufficient
661 to delineate species boundaries in *Symbiodinium*. *Molecular ecology*, **25**, 2719-2723.
- 662 Jakubowska M, Normant M (2015) Metabolic rate and activity of blue mussel *Mytilus edulis*
663 *trossulus* under short-term exposure to carbon dioxide-induced water acidification and
664 oxygen deficiency. *Marine and Freshwater Behaviour and Physiology*, **48**, 25-39.
- 665 Jansson A, Norkko J, Dupont S, Norkko A (2015) Growth and survival in a changing
666 environment: Combined effects of moderate hypoxia and low pH on juvenile bivalve
667 *Macoma balthica*. *Journal of Sea Research*, **102**, 41-47.
- 668 Jarrold MD, Calosi P, Verberk WCEP, Rastrick SPS, Atfield A, Spicer JI (2013) Physiological
669 plasticity preserves the metabolic relationship of the intertidal non-calcifying anthozoan-
670 *Symbiodinium* symbiosis under ocean acidification. *Journal of Experimental Biology and
671 Ecology*, **449**, 200-206.

- 672 Kim T, Barry J, Micheli F (2013) The effects of intermittent exposure to low-pH and low-oxygen
673 conditions on survival and growth of juvenile red abalone. *Biogeosciences*, **10**, 7255-7262.
- 674 Klein SG, Pitt KA, Carroll AR (2016) Surviving but not thriving: inconsistent responses of
675 zooxanthellate jellyfish polyps to ocean warming and future UV-B scenarios. *Scientific*
676 *Reports*, **6**, 28859.
- 677 Köhler-Rink S, Kühl M (2000) Microsensor studies of photosynthesis and respiration in larger
678 symbiotic foraminifera. I The physico-chemical microenvironment of *Marginopora*
679 *vertebralis*, *Amphistegina lobifera* and *Amphisorus hemprichii*. *Marine Biology*, **137**, 473-
680 486.
- 681 Kroeker KJ, Kordas RL, Crim RN, Singh GG (2010) Meta-analysis reveals negative yet variable
682 effects of ocean acidification on marine organisms. *Ecology Letters*, **13**, 1419-1434.
- 683 Lajeunesse TC (2005) “Species” radiations of symbiotic dinoflagellates in the Atlantic and Indo-
684 Pacific since the Miocene-Pliocene transition. *Molecular Biology and Evolution*, **22**, 570-
685 581.
- 686 Landry CA, Steele SL, Manning S, Cheek AO (2007) Long term hypoxia suppresses reproductive
687 capacity in the estuarine fish, *Fundulus grandis*. *Comparative Biochemistry and*
688 *Physiology Part A: Molecular & Integrative Physiology*, **148**, 317-323.
- 689 Larkum AW, Koch E-M, Kühl M (2003) Diffusive boundary layers and photosynthesis of the
690 epilithic algal community of coral reefs. *Marine Biology*, **142**, 1073-1082.
- 691 Laurent J, Tambutté S, Tambutté É, Allemand D, Venn A (2013) The influence of photosynthesis
692 on host intracellular pH in scleractinian corals. *Journal of Experimental Biology*, **216**,
693 1398-1404.
- 694 Laurent J, Venn A, Tambutté É, Ganot P, Allemand D, Tambutté S (2014) Regulation of
695 intracellular pH in cnidarians: response to acidosis in *Anemonia viridis*. *Febs Journal*, **281**,
696 683-695.
- 697 Lemloh M-L, Fromont J, Brümmer F, Usher KM (2009) Diversity and abundance of
698 photosynthetic sponges in temperate Western Australia. *BMC ecology*, **9**, 1-13.
- 699 Levin RA, Suggett DJ, Nitschke MR, Oppen MJ, Steinberg PD (2017) Expanding the
700 *Symbiodinium* (Dinophyceae, Suessiales) toolkit through protoplast technology. *Journal of*
701 *Eukaryotic Microbiology*, doi:10.1111/jeu.12393
- 702 Lewis CL, Coffroth MA (2004) The acquisition of exogenous algal symbionts by an octocoral
703 after bleaching. *Science*, **304**, 1490-1492.

- 704 Lewis E, Wallace D, Allison LJ (1998) Program Developed for CO₂ System Calculations.
705 ORNL/CDIAC-105. Carbon dioxide Information Analysis Center, Oak Ridge National
706 Laboratory, US Department of Energy, Oak Ridge, TN.
- 707 Little AF, Van Oppen MJ, Willis BL (2004) Flexibility in algal endosymbioses shapes growth in
708 reef corals. *Science*, **304**, 1492-1494.
- 709 Malcolm JM, Brown WI (1977) Zooxanthellae-produced O₂ promotes sea anemone expansion
710 and eliminates oxygen debt under environmental hypoxia. *Journal of Experimental*
711 *Zoology*, **201**, 149-155.
- 712 Mellas RE, McIlroy SE, Fitt WK, Coffroth MA (2014) Variation in symbiont uptake in the early
713 ontogeny of the upside-down jellyfish, *Cassiopea* spp. *Journal of Experimental Marine*
714 *Biology and Ecology*, **459**, 38-44.
- 715 Melzner F, Thomsen J, Koeve W *et al.* (2012) Future ocean acidification will be amplified by
716 hypoxia in coastal habitats. *Marine Biology*, 1-14.
- 717 Park K, Kim C-K, Schroeder WW (2007) Temporal variability in summertime bottom hypoxia in
718 shallow areas of Mobile Bay, Alabama. *Estuaries and Coasts*, **30**, 54-65.
- 719 Przeslawski R, Byrne M, Mellin C (2015) A review and meta-analysis of the effects of multiple
720 abiotic stressors on marine embryos and larvae. *Global Change Biology*, **21**, 2122-2140.
- 721 Putnam HM, Stat M, Pochon X, Gates RD (2012) Endosymbiotic flexibility associates with
722 environmental sensitivity in scleractinian corals. *Proceedings of the Royal Society of*
723 *London B: Biological Sciences*, doi:10.1098/rspb.2012.1454.
- 724 Rabalais N, Diaz R, Levin L, Turner R, Gilbert D, Zhang J (2010) Dynamics and distribution of
725 natural and human-caused hypoxia. *Biogeosciences*, **7**, 585-619.
- 726 Reece JB, Urry LA, Cain ML, Wasserman SA, Minorsky PV, Jackson RB (2015) Campbell
727 *Biology: Photosynthesis*. pp 199-201. Pearson, Boston.
- 728 Regnault M, Aldrich JC (1988) Short-term effect of hypoxia on ammonia excretion and
729 respiration rates in the crab *Carcinus maenas*. *Marine & Freshwater Behaviour &*
730 *Physiology*, **13**, 257-271.
- 731 Revsbech NP, Kühl M, Cohen Y, Dalsgaard T, Jørgensen B (1995) Microenvironment and
732 photosynthesis of zooxanthellae in scleractinian corals studied with microsensors for O₂,
733 pH and light. *Marine Ecology Progress Series*, **117**, 159-172.
- 734 Ritchie RJ (2006) Consistent sets of spectrophotometric chlorophyll equations for acetone,
735 methanol and ethanol solvents. *Photosynthesis Research*, **89**, 27-41.
- 736 Roleda MY, Morris JN, McGraw CM, Hurd CL (2012) Ocean acidification and seaweed
737 reproduction: increased CO₂ ameliorates the negative effect of lowered pH on meiospore

738 germination in the giant kelp *Macrocystis pyrifera* (Laminariales, Phaeophyceae). *Global*
739 *Change Biology*, **18**, 854-864.

740 Sachs JL, Wilcox TP (2006) A shift to parasitism in the jellyfish symbiont *Symbiodinium*
741 *microadriaticum*. *Proceedings of the Royal Society of London B: Biological Sciences*, **273**,
742 425-429.

743 Santos SR, Taylor DJ, Kinzie Iii RA, Hidaka M, Sakai K, Coffroth MA (2002) Molecular
744 phylogeny of symbiotic dinoflagellates inferred from partial chloroplast large subunit
745 (23S)-rDNA sequences. *Molecular phylogenetics and evolution*, **23**, 97-111.

746 Schmidtko S, Stramma L, Visbeck M (2017) Decline in global oceanic oxygen content during the
747 past five decades. *Nature*, **542**, 335-339.

748 Seutin G, White BN, Boag PT (1991) Preservation of avian blood and tissue samples for DNA
749 analyses. *Canadian Journal of Zoology*, **69**, 82-90.

750 SPSS (Released 2013) IBM SPSS Statistics for Windows. *Version 22.0, Armonk, NY: IBM Corp.*

751 Stat M, Bird CE, Pochon X *et al.* (2011) Variation in *Symbiodinium* ITS2 sequence assemblages
752 among coral colonies. *PLOS one*, **6**, e15854.

753 Steckbauer A, Ramajo L, Hendriks IE, Fernandez M, Lagos NA, Prado L, Duarte CM (2015)
754 Synergistic effects of hypoxia and increasing CO₂ on benthic invertebrates of the central
755 Chilean coast. *Frontiers in Marine Science*, **2**, 49.

756 Steindler L, Beer S, Ilan M (2002) Photosymbiosis in intertidal and subtidal tropical sponges.
757 *SYMBIOSIS-REHOVOT-*, **33**, 263-274.

758 Suggett DJ, Dong LF, Lawson T, Lawrenz E, Torres L, Smith DJ (2013) Light availability
759 determines susceptibility of reef building corals to ocean acidification. *Coral reefs*, **32**,
760 327-337.

761 Suggett DJ, Hall-Spencer JM, Rodolfo-Metalpa R *et al.* (2012) Sea anemones may thrive in a high
762 CO₂ world. *Global Change Biology*, **18**, 3015-3025.

763 Sui Y, Kong H, Shang Y *et al.* (2016) Effects of short-term hypoxia and seawater acidification on
764 hemocyte responses of the mussel *Mytilus coruscus*. *Marine Pollution Bulletin*, **108**, 46-
765 52.

766 Thornhill DJ, Daniel MW, Lajeunesse TC, Schmidt GW, Fitt WK (2006) Natural infections of
767 aposymbiotic *Cassiopea xamachana* scyphistomae from environmental pools of
768 *Symbiodinium*. *Journal of Experimental Marine Biology and Ecology*, **338**, 50-56.

769 Tonetta D, Fontes MLS, Petrucio MM (2014) Determining the high variability of pCO₂ and pO₂
770 in the littoral zone of a subtropical coastal lake. *Acta Limnologica Brasiliensia*, **26**, 288-
771 295.

- 772 Tonk L, Bongaerts P, Sampayo EM, Hoegh-Guldberg O (2013) SymbioGBR: a web-based
773 database of *Symbiodinium* associated with cnidarian hosts on the Great Barrier Reef. *BMC*
774 *Ecology*, **13**, 1.
- 775 Towanda T, Thuesen EV (2012) Prolonged exposure to elevated CO₂ promotes growth of the
776 algal symbiont *Symbiodinium muscatinei* in the intertidal sea anemone *Anthopleura*
777 *elegantissima*. *Biology Open*, **1**, 615-621.
- 778 Tyler RM, Brady DC, Targett TE (2009) Temporal and spatial dynamics of diel-cycling hypoxia
779 in estuarine tributaries. *Estuaries and Coasts*, **32**, 123-145.
- 780 Uthicke S, Fabricius KE (2012) Productivity gains do not compensate for reduced calcification
781 under near-future ocean acidification in the photosynthetic benthic foraminifer species
782 *Marginopora vertebralis*. *Global Change Biology*, **18**, 2781-2791.
- 783 Vaquer-Sunyer R, Duarte CM (2008) Thresholds of hypoxia for marine biodiversity. *Proceedings*
784 *of the National Academy of Sciences*, **105**, 15452-15457.
- 785 Venn A, Ta Venn A, Tambutté E, Lotto S, Zoccola D, Allemand D, Tambutté S (2009) Imaging
786 intracellular pH in a reef coral and symbiotic anemone. *Proceedings of the National*
787 *Academy of Sciences*, **106**, 16574-16579.
- 788 Ventura P, Jarrold MD, Merle P-L *et al.* (2016) Resilience to ocean acidification: decreased
789 carbonic anhydrase activity in sea anemones under high pCO₂ conditions. *Marine Ecology*
790 *Progress Series*, **559**, 257-263.
- 791 Wang W, Widdows J (1991) Physiological responses of mussel larvae *Mytilus edulis* to
792 environmental hypoxia and anoxia. *Marine Ecology Progress Series*, **70**, 223-236.
- 793 Wham DC, Lajeunesse TC (2016) *Symbiodinium* population genetics: testing for species
794 boundaries and analysing samples with mixed genotypes. *Molecular Ecology*,
795 doi:10.1111/mec.13623
- 796 Woulds C, Cowie GL, Levin LA *et al.* (2007) Oxygen as a control on sea floor biological
797 communities and their roles in sedimentary carbon cycling. *Limnology and Oceanography*,
798 **52**, 1698.
- 799 Young CS, Gobler CJ (2016) Ocean acidification accelerates the growth of two bloom-forming
800 macroalgae. *PLOS one*, **11**, e0155152.

801 **Tables**

802 **Table 1** Summary of results for a LMMs analysis comparing the number of polyps between
803 treatments at day 22 of the experiment. Df= Degrees of freedom. *P* values in bold are statistically
804 significant (*P* < 0.05). BIC (Bayesian Information Criterion) =85.204, and AIC (Akaike's

805 Information Criterion) =84.026. For all sources of variation numerator df =1 and denominator df
 806 =24.

Source	F	Sig.
Symbiont	106.778	< 0.001
Oxygen	49.00	< 0.001
pH	0.111	0.742

Source	Numerator df	Denominator df	F	Sig.
Symbiont × Oxygen	1.000	0.327		
Symbiont × pH	25.000	< 0.001		
Oxygen × pH	0.111	0.742		
Symbiont × Oxygen × pH	0.111	0.742		

807

808 **Table 2** Summary of results for a LMMs analysis comparing day and night pH microelectrode
 809 profiles between treatments at Day 22 of the experiment. The model-of-best-fit was AR(1), BIC
 810 (Bayesian Information Criterion) = -1268.564, AIC (Akaike's Information Criterion) = -1276.198.
 811 Note, the factor Oxygen was included in preliminary analysis but removed due to non-
 812 significance for all terms ($P>0.05$) and LMMs analysis was re-run. Df= Degrees of freedom
 813 (numerator, denominator). P values in bold are statistically significant ($P < 0.05$).

Source of variation	Df	MS	F	P
Day/ Night	1	89.181	309.977	<0.001
Symbiont	1	89.181	120.335	<0.001
pH	1	89.181	3810.007	<0.001
Day/ Night × Symbiont	1	89.181	98.877	<0.001
Day/ Night × pH	1	89.181	7.747	0.007
Symbiont × pH	1	89.181	17.477	<0.001
Day/ Night × Symbiont × pH	1	89.181	10.509	0.002
Distance	5	239.171	45.393	<0.001
Day/ Night × Distance	5	239.171	19.691	<0.001
Symbiont × Distance	5	239.171	36.126	<0.001
pH × Distance	5	239.171	12.215	<0.001
Day/ Night × Symbiont × Distance	5	239.171	32.763	<0.001
Day/ Night × pH × Distance	5	239.171	13.975	<0.001
Symbiont × pH × Distance	5	239.171	5.520	<0.001
Day/ Night × Symbiont × pH × Distance	5	239.171	4.628	<0.001

Table 3

Summary of results for three LMMs comparing *Symbiodinium* density (cell polyp⁻¹) and *chl a* cell⁻¹ (pg) between

827 treatments of symbiotic polyps at Day 22 of the experiment. Df = degrees of freedom. BIC= 828 Bayesian Information Criterion and AIC = Akaike's Information Criterion. *P* values in bold are 829 statistically significant (*P* < 0.05). For all sources of variation numerator df =1 and denominator 830 df =24.

Variable	<i>Symbiodinium</i> polyp ⁻¹	<i>Chla</i> cell ⁻¹
Transformation	Ln	None
Information Criterion	BIC= 10.359 AIC= 9.874	BIC= 59.840 AIC= 59.355
Source of variation	<i>P</i>	<i>P</i>
pH	0.131	0.035
Oxygen	<i>F</i> =2.633 0.124	<i>F</i> =5.623 0.419
pH × Oxygen	<i>F</i> =2.720 0.047	<i>F</i> =0.700 0.002

Figure 1

Mean ±1SE number of polyps recorded at

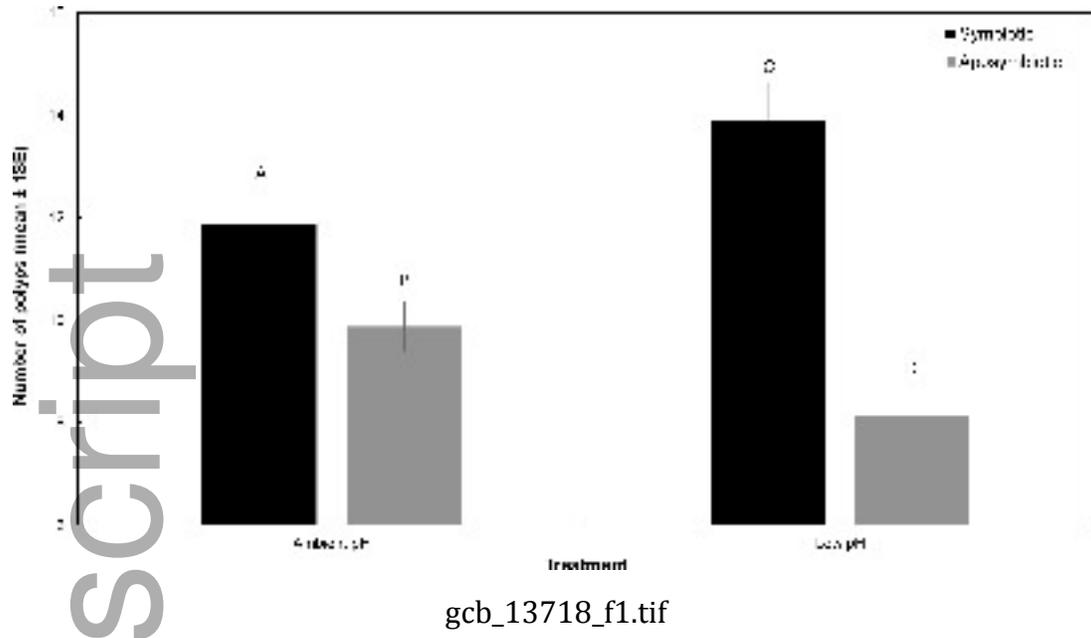
840 Day 22 of the experiment. Letters above error bars indicate similarities (e.g. AA) or differences 841 (e.g. AB) between treatments, as determined by estimated marginal means.

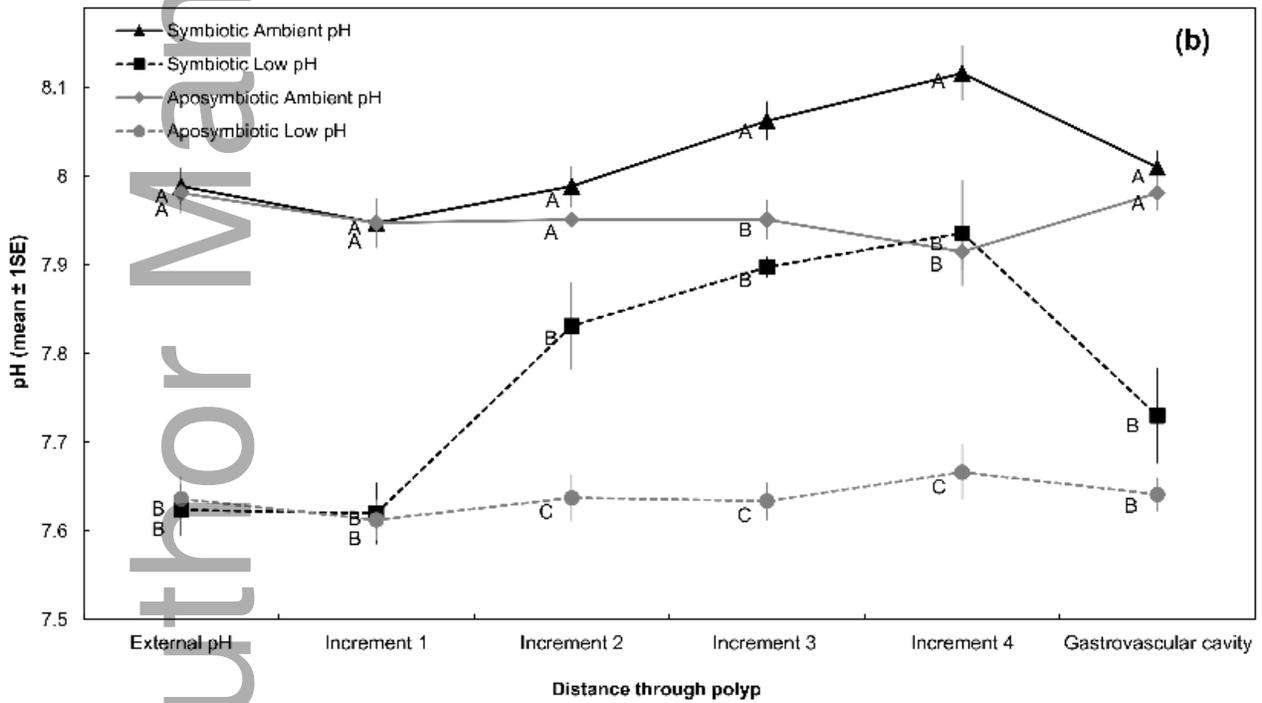
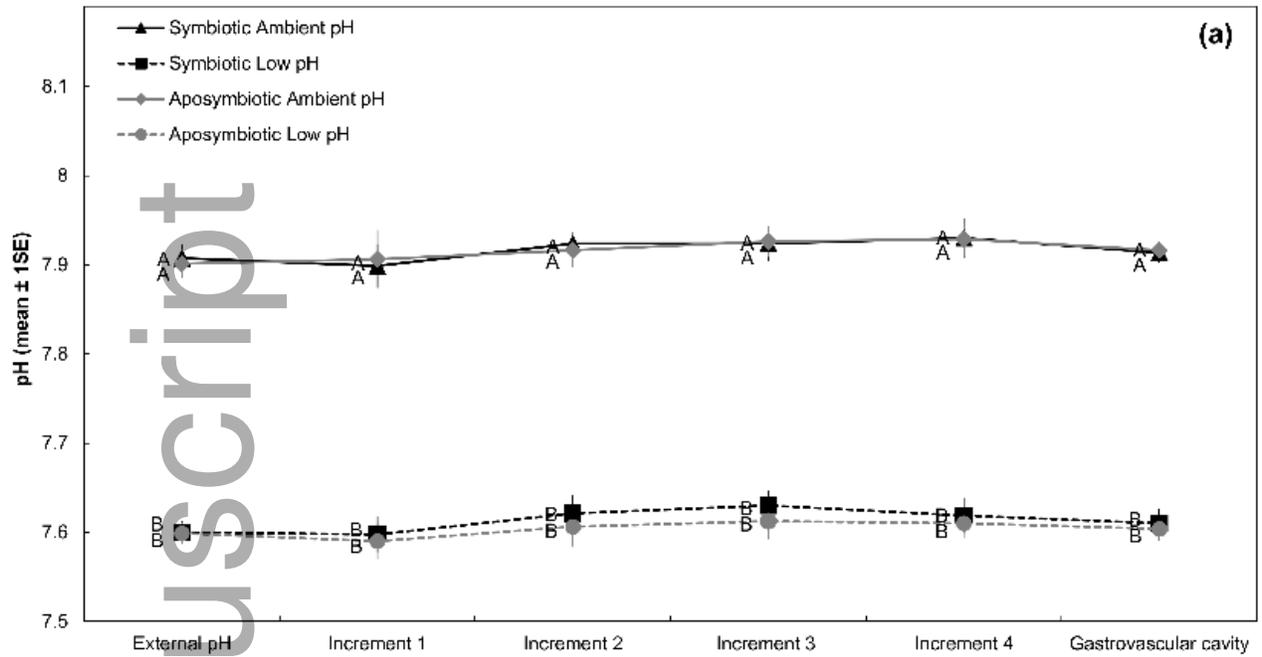
842
843 **Figure 2** Mean (±1SE) pH measurements taken through polyp walls taken during the night (a) and 844 day (b). Letters next to data points indicate similarities (e.g. AA) or differences (e.g. AB) between 845 treatments, as determined by estimated marginal means.

846

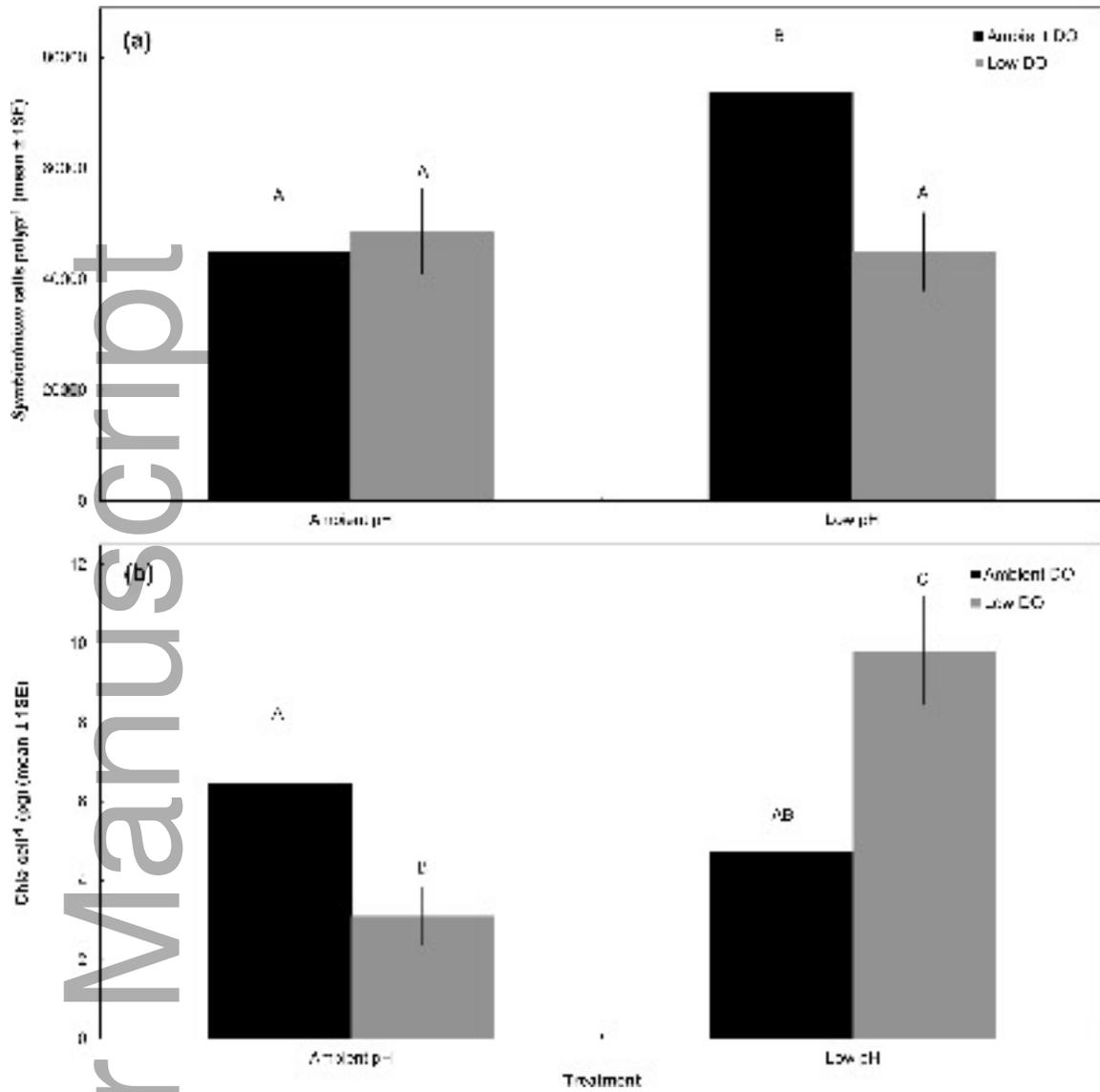
847 **Figure 3** Mean ($\pm 1SE$) *Symbiodinium* cells polyp⁻¹ (a) Chla *Symbiodinium* cell⁻¹ (pg) (b) at Day
848 22 of the experiment. Letters above error bars indicate similarities (e.g. AA) or differences (e.g.
849 AB) between treatments, as determined by estimated marginal means.

Author Manuscript





gcb_13718_f2.tif



gcb_13718_f3.tif