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Electrochemical and hydraulic analysis of thin-film composite and cellulose triacetate membranes for seawater electrolysis applications

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- 2 Submitted to: Journal of Membrane Science
- 3 Electrochemical and Hydraulic Analysis of Thin-Film Composite and 4 **Cellulose Triacetate Membranes for Seawater Electrolysis Applications** 5 6 7 Rachel Taylor¹, Le Shi², Xuechen Zhou², Ruggero Rossi², Cristian Picioreanu³, and 8 Bruce E. Logan^{1,2} 9 ¹Department of Chemical Engineering, The Pennsylvania State University, University Park, PA, USA. 10 ²Department of Civil and Environmental Engineering, The Pennsylvania State University, University Park, PA, 11 USA. 12 ³Water Desalination and Reuse Center (WDRC), Biological and Environmental Science and Engineering 13 Division (BESE), King Abdullah University of Science and Technology (KAUST), Thuwal, Saudi Arabia *Corresponding author. Email: blogan@psu.edu; Tel.: +1-814-863-7908 14 15 16

18 Abstract

19

20 Thin film, reverse osmosis (RO) membranes can be used in saltwater electrolysis to minimize chlorine ion

crossover and chlorine gas generation without significantly increasing the potential requirement for electrolysis in comparison to a cation exchange membrane (CEM). Optimizing membrane performance

requires a better understanding of membrane properties that impact electrical resistances and ion retention.

24 Twelve RO membranes, one nanofiltration (NF) membrane, and one cellulose triacetate forward osmosis

25 (FO) membrane were examined for their electrical resistances under conditions typically used for CEMs.

26 Resistances measured at low current densities (0.07 to 0.3 mA/cm²) varied between different membranes

- by over an order of magnitude in 1 M NaCl at neutral pH, from 6.1 ± 0.1 W cm² to 70 ± 30 W cm². There
- 28 was no significant correlation between membrane resistance and applied potential during saltwater
- electrolysis at 20 mA/cm² (p=0.44), or between membrane resistance and water permeability (p=0.35).
 These results indicate traditional CEM resistance characterization methods do not predict RO membrane
- 30 These results indicate traditional CEM resistance characterization methods do not predict RO membrane 31 electrolysis performance because proton and hydroxide transport, which is important during electrolysis
- 32 when large pH gradients develop, must be considered separately from salt ion and water molecule transport
- through size selective RO, NF, and FO membranes during water electrolysis.
- 34

35 Keywords: Seawater Electrolysis, Thin Film Composite, Cellulose Triacetate, Electrical Resistance, Ion

36 Transport

37 1. Introduction

38

39 Ultra-pure water is currently required for water electrolysis to produce carbon neutral (green) hydrogen 40 through proton exchange membrane (PEM) electrolysis [1-3]. Research into using less purified water is gaining momentum because low-grade waters such as brackish water or seawater is more globally 41 42 accessible, while coastal regions, where seawater is available, typically have better access to renewable 43 energy sources compared to onshore locations. Additionally, producing highly deionized water makes the 44 overall process more complex and expensive [4, 5]. Direct seawater electrolysis has been difficult to 45 implement due to the presence of chloride ions in seawater which react at the anode to form undesirable 46 species such as chlorine and derivative species (e.g. hypochlorite and hypochlorous acid), which can 47 damage electrolyzer components [6]. While there has been extensive research into developing novel 48 catalysts that preferentially evolve oxygen over chlorine, an alternative approach is using thin-film 49 composite reverse osmosis (RO) membranes with a contained analyte and a seawater catholyte [7-12]. An 50 anolyte that contains fully oxidized salt species such as perchlorate (NaClO₄) can be used as an inert 51 electrolyte, while the membrane prevents chloride ion transfer to the anode and thus its oxidation to chlorine 52 gas and other species. Traditional cation exchange membranes (CEMs) cannot be used under these 53 conditions because they allow too much chloride ion leakage from the catholyte to the anolyte [1]. RO 54 membranes, have an additional advantage of being substantially less expensive than CEMs [12].

55

56 Reverse osmosis, forward osmosis (FO), and nanofiltration (NF) membranes are three types of size-57 selective membranes that are being studied as alternatives to ion exchange membranes. RO membranes 58 have the greatest selectivity for preventing transport of salt species, but they must facilitate transport of 59 small ions, such as protons/hydronium and hydroxide, to maintain the high current densities needed in water 60 electrolyzers [13-16]. RO, NF, and FO membranes are size selective due to their highly dense active layers [17-19]. Thin film composite RO and NF membranes are typically composed of three layers, including a 61 62 dense, polyamide active layer, which can range between 20-150 nm thick for NF membranes and 100-200 63 nm thick for RO membranes [20]. The active layer is bound to a polysulfone support layer (~50 mm) which 64 is used to connect the thin, fragile active layer to a thick ($\sim 100 \text{ mm}$) polyester web backbone. The polyester web backbone maintains the mechanical properties of the membranes in large hydraulic pressure gradients 65 [14, 21, 22]. While FO membranes also have a size-selective active layer, they differ in that a porous support 66 67 layer is not required to have mechanical strength to withstand higher pressures because water transport 68 across FO membranes is only driven by an osmotic pressure gradient [23]. CEMs are usually homogenous 69 block co-polymers, charged to selectively transport all positive species in a solution. This does not allow 70 selectivity between salt ions, which must be contained in their respective compartments, and protons, which are the electrochemically active species and the preferable charge carrier [24]. Filtration membranes allow 71 72 for selective ion transport, but the electrochemical properties of RO, NF, and FO membranes have yet to 73 be broadly studied for applications in electro-driven separations such as saltwater electrolysis.

74

75 In electro-driven processes, membrane electrical resistance is one of several factors used to compare 76 electrolyzer performance [24-26]. Electrical resistance is a measure of a membrane's ability to transport 77 electrical charge in the form of ions across it, and therefore ion exchange membranes with higher resistances 78 increase energy consumption for water electrolysis [25, 26]. The electrical resistance of the RO membrane 79 active layer has been measured using electrochemical impedance spectroscopy and related to salt 80 permeability during desalination tests with a large water flux through the membrane. Few studies have examined the resistance of the RO membrane active layer and supporting layers together, which is the 81 82 critical property of these membranes in electrochemical applications [15, 27-29]. In the first study of using 83 RO membranes for water electrolyzers, Shi et al. showed that one RO membrane had an electrical resistance 84 comparable to CEMs, while another had a much larger electrical resistance. The membrane with a resistance comparable to the CEM was used to electrolyze saltwater at an applied potential similar to that used with 85 86 the CEM, while using the membrane with the higher resistance required a higher applied potential during 87 electrolysis, but the reason for this difference in performance between the two RO membranes was not presented [12]. Because only two RO membranes were used, it was not possible to examine if other
membrane properties, such as water permeability, could be used to predict performance in a water
electrolyzer [12].

91

92 The purpose of this study was to determine if resistances of different size-selective filtration membranes, 93 measured under conditions typically used to characterize resistances of CEMs, could be correlated with 94 water electrolyzer performance at high current densities. At high current densities, large pH gradients 95 develop, and water ions account for a larger percentage of charge transport across the membrane, while in 96 low current density resistance tests, salt ions primarily transport charge across the membrane. Electrical 97 resistances were measured at a low current density in neutral pH for twelve RO membranes, one NF 98 membrane, and one FO membrane. Three RO membranes with varying electrical resistances, and the FO 99 and NF membranes were selected for further characterization in electrolyzer and permeability tests to 100 examine whether water flux could be used to better understand membrane performance in water 101 electrolyzers. Therefore, membrane overpotential, salt ion crossover during electrolysis, and membrane permeability were examined to see if any of these properties could be correlated to membrane resistance or 102 be used to determine optimal characteristics of TFC membranes for green hydrogen production from 103 104 saltwater.

106 **2.** Experimental

107

105

108 2.1 Membranes

109 110 The following flat-sheet RO membranes were used (membrane abbreviations in parentheses) based on the manufacturer and intended application of either brackish water (BW) or seawater desalination (SW) 111 desalination: Trisep Membrane ACM5 [RO1 (BW)]; DuPont membranes BWXLE [RO2 (BW)], 112 BW30XLE [RO4 (BW)], SWXLE [RO5 (SW)], BW30 [RO6 (BW)], BW30LE [RO7 (BW)]; Hydranautics 113 114 membranes SWC4 [RO3 (SW)], SWC5 [RO8 (SW)]; Toray membranes 73AC [RO9 (BW)], 73HA [RO10 115 (BW)]; GE Suez membranes GE AK [RO11 (BW)], GE AG [RO12 (BW)]. The nanofiltration membrane was DuPont NF270 (NF), and the forward osmosis membrane was Fluid Technology Solutions CTA FO 116 (FO). A variety of RO membranes with different rejections, manufacturers, and advertised energy 117 118 requirements were studied to understand how much RO membrane electrochemical performance can vary. 119 An outside micrometer (Mitutoyo Kawasaki, Japan) was used to measure the total thickness of dry 120 membranes.

121

122 2.2 Resistance Measurements123

124 The membrane resistances were determined using a four-electrode direct current method typically 125 employed for CEMs [12, 30, 31]. The electrodes and membrane were submerged in 60 mL of electrolyte solution, in a cylindrical chamber with a cross-sectional area of 7 cm². The membrane was positioned in 126 the middle of the chamber, 5 cm away from the anode and cathode. A scheme and picture of the 127 experimental set-up is in the supporting information (Supporting Information, Figure S1 and Figure S2). 128 Platinum coated titanium mesh electrodes were used as the anode and cathode, placed 10 cm apart. Two 129 130 Ag/AgCl (3M NaCl) reference electrodes (BASi West Lafayette, IN) with Luggin capillaries were placed on either side of the membrane. The Luggin capillaries minimized the ohmic drop between the reference 131 electrodes [32]. The exposed membrane area, 7 cm², was the same as the cross section of the cylindrical 132 133 chamber.

134

135 A potentiostat (Biologic VMP3) was used to obtain linear sweep voltammetry (LSV) data from 0V to 3.5V

- 136 at a scan rate of 5 mV/s. From this data, the ohmic region was determined, and current densities were 137 selected for membrane electric resistance measurements. In the ohmic region, the applied potential, U(V),
- changes linearly with current, I(A), and the proportionality constant between the two is the ohmic resistance

139 R (W), consistent with Ohm's law, U = IR [33]. Measuring the potential across a membrane at current 140 densities in the ohmic region will yield the ohmic resistance of the membrane as the slope of the potential 141 vs. current data. Eight current densities were selected that ranged from 0.07 to 0.3 mA/cm² in 142 chronopotentiometry (CP) tests based on 30 s intervals. An example of how resistance is calculated from 143 chronopotentiometry data is provided in the SI.

144

Membrane resistance (R_{mem}) was calculated from measurements made in the absence and presence of the membrane, based on $R_{mem} = R_{sol + mem} - R_{sol}$, where R_{sol} (the solution resistance measured without membrane) was subtracted from the total resistance with the membrane present, $R_{sol+mem}$. The area resistance (W cm²) of the membrane was calculated using the exposed membrane area. Measurements were repeated 3 times, using a fresh piece of membrane each time. The resistances for each membrane were measured in both 0.6 M (~3.5 wt%) and 1 M (~5.5 wt%) NaCl. The membranes were equilibrated in the electrolyte for three days prior to the experiment.

- 152
- 153 2.3 Water Flux Measurements154

The water flux through the membrane was measured using a high-pressure dead-end cell (HP 4759,
 Sterlitech, Auburn, WA). Deionized water was used at an applied pressure of 34.4 bar. The mass of
 permeate was measured over time using a scale and used to calculate the flux normalized to pressure (L m⁻²
 h⁻¹ bar ⁻¹ or, shortly, LMH/bar). The exposed membrane area was 14.6 cm².

159

160 2.4 Water Electrolyzer Experiments161

162 The applied cell potentials required for saltwater electrolysis with an asymmetric anolyte and catholyte 163 contained by each membrane was compared at a constant current. A zero-gap electrolyzer (Scribner, North Carolina) configuration was used to investigate performance in a conventional proton exchange membrane 164 165 (PEM) electrolyzer [34, 35]. In the zero-gap electrolyzer, the membrane is placed directly next to the anode 166 and cathode to minimize ohmic resistance. The anode and cathode are placed adjacent to serpentine flow 167 channels carrying the anolyte and catholyte at a rate of 15 mL/s. A platinized titanium flow field is used for 168 the anode and a graphite flow field is used for the cathode. Gold plated copper current collectors are attached 169 to the Pt-Ti and graphite flow fields, and all components are contained between two exterior anodized 170 aluminum end plates held together with bolts and washers.

171

172 A two-electrode set-up was used to perform the experiment and measure the applied potential required for 173 saltwater electrolysis. Carbon cloth electrodes (4 cm²) coated with 10% Pt/C catalyst were used as the anode and cathode. Experiments were quick to avoid damaging the electrodes. The anolyte was 1 M NaClO₄ and 174 175 catholyte was 1 M NaCl with the electrolytes (each 200 mL) recirculated through the flow cell at a rate of 176 15 mL/s. The exposed membrane area was 5 cm². Chronopotentiometry (CP) was used at a constant current 177 density of 20 mA/cm² using a potentiostat (VMP3, Bio-logic). Experiments were run for three hours, and 178 the applied potentials were compared at the end of the three hours. Each experiment was conducted twice 179 for each membrane type, using a fresh piece of membrane and fresh electrodes each time (n=2).

- 180
- 181 2.5 Ion Crossover Experiments

182

A two-electrode set-up was used to measure ion transport across the membranes, as previously described [12]. The anode and cathode were both carbon paper coated with a 10% platinum/carbon catalyst and had areas of 1.68 cm². The catholyte was 30 mL of 1 M KCl, and the anolyte was 30 mL of 1 M NaClO₄. K⁺ was used in the catholyte so the Na⁺ crossover could be studied in the direction of the electric field. The anolyte and catholyte were added into the cylindrical cell, and chronopotentiometry was used to apply a current density of 10 mA/cm² for one hour to facilitate electrolysis. After the hour, the anolyte and catholyte were collected, the pH of both were measured, and ion chromatography (Dionex, Thermo Fisher Waltham, 190 MA) was used to measure the concentration of ions that crossed over into each compartment, (Na⁺ and 191 ClO_4^- in the catholyte, or K⁺ and Cl⁻ in the anolyte). The membrane had an exposed area of 7 cm².

192

194

- **193 3. Results and Discussion**
- 195 *3.1 Membrane Resistances*

196 197 Membrane resistances varied by an order of magnitude with no apparent correlation of the type of RO 198 membrane (BW or SW) with resistance. among all the studied membranes. In 1 M NaCl, the RO1 (BW) 199 membrane had the lowest resistance of 6.1 ± 0.1 W cm², an order of magnitude lower than the RO9 (BW) 200 membrane, with the highest resistance of 70 ± 30 W cm² (Figure 1). BWRO and SWRO membranes had 201 similar resistances, despite their different reported salt rejections and permeabilities [36]. The nanofiltration 202 membrane has a NaCl rejection much lower than all the RO membranes (50% compared to 98.5-99.8%) 203 according to manufacturer data, and one of the lowest measured resistances in 1 M NaCl, 14 ± 1.7 W cm² [37]. A BWRO membrane with a reported salt rejection of 98.5% had an even lower resistance (RO1), 204 205 suggesting that the same properties that control one directional ion transport during water filtration do not 206 control two directional ion transport across membranes in potential gradient [38, 39]. The FO membrane 207 had a resistance of 24 ± 5.6 W cm², which was in the midrange in comparison to the other membrane 208 resistances.

209

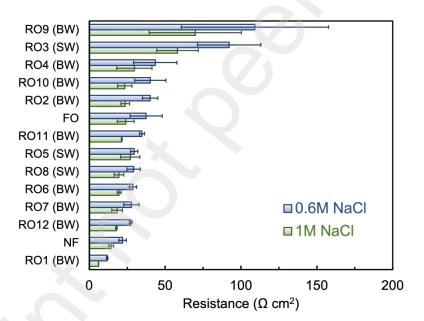


Figure 1. Resistances of reverse osmosis, nanofiltration, and forward osmosis membranes measured in either 0.6 M or 1 M NaCl using the four-electrode direct current method.

The membrane resistance increased as the solution concentration decreased. On average, the resistances of the membranes increased by $35 \pm 8\%$ from the 1 M to 0.6 M NaCl solution, consistent with previous studies showing that membrane electrical resistance is a function of electrolyte concentration [31, 40].

213

The total membrane thickness did not have any correlation with the membrane resistance (p = 0.5, Supporting Information, Figure S5). For ion exchange membranes, total thickness is directly related to membrane resistance because ion exchange membranes are homogenous block co-polymers [41]. The asymmetric layers in filtration membranes have different transport properties that may complicate identifying simple correlations between membrane thickness and resistance [17].

219

220 3.2 Water permeabilities

221

Membrane permeabilities ranged from 0.7 L m⁻² h⁻¹ bar⁻¹ (RO3) to 14.6 L m⁻² h⁻¹ bar⁻¹ (NF) (Figure 2a). 222 223 There was a general increase in resistance with less water permeability, but there was no significant trend $(y = 20.97e^{-0.04x}, R^2 = 0.67, p = 0.35)$ (Figure 2b). The permeabilities of three RO membranes with 224 varying electrical resistances were measured, as well as the nanofiltration membrane due to its different 225 composition. The FO membrane permeability was not measured because the membrane is only used with 226 an osmotic pressure gradient and thus it cannot withstand the high hydraulic pressure used in the test. The 227 228 nanofiltration membrane had the highest measured permeability, at 14.6 ± 0.9 L m⁻² h⁻¹ bar⁻¹, which is similar to that reported in literature (Figure 2a) [42, 43]. NF membranes have larger pore sizes (0.5 - 2)229 nm) than RO membranes (0.2-1 nm) because they are typically used to filter organic compounds and 230 soften surface and ground water by separating divalent ions [38, 44-46]. Therefore, NF membranes have 231 higher water and NaCl permeability than RO membranes. The SWRO membrane RO3, which had the 232 highest resistance (Figure 1), had the lowest water permeability, 0.7 ± 0.1 L m⁻² h⁻¹ bar⁻¹. SWRO 233 234 membranes are typically used in higher pressure gradients with higher concentration salt solutions, so they have higher salt rejections and lower permeabilities to achieve the desired separation [47]. When 235 236 only the RO membrane permeabilities were plotted against their electric resistance, the fitted curve had R²

- 237 = 1 but the trend was not significant (p = 0.08) possible due to the limited number of data points.
- 238

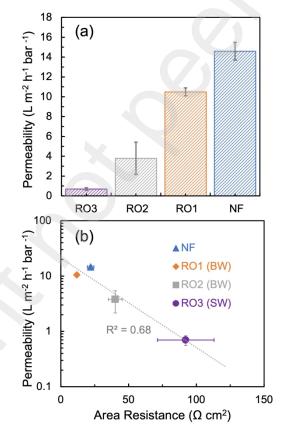


Figure 2. (a) Water permeabilities of three reverse osmosis membranes and one nanofiltration membrane measured at an applied pressure of 34.4 bar. (b) Water permeabilities as a function of measured area resistance.



3.3 Membrane performance in electrolysis

242 Membranes with higher resistances required higher voltage during saltwater electrolysis, but statistically, 243 membrane resistance and applied potential required for saltwater electrolysis could not be shown to be 244 significantly correlated (p = 0.44). Electrolyzing saltwater with the FO membrane required the smallest 245 applied potential of 2.8 ± 0 V, although this membrane did not have the smallest measured resistance (Figure 3a). The membrane with the smallest measured resistance, RO1 (BW), had a similar applied 246 potential of 2.9 ± 0.1 V. Electrolysis with the RO2 (BW) membrane required the highest applied potential 247 248 of 4.1 ± 0.1 V, while the RO3 (SW) membrane, which had the largest measured membrane resistance, had 249 an applied potential of 3.8 ± 0.1 V. The membrane resistances in 0.6 M NaCl were plotted against the 250 applied potential required for electrolysis using each membrane. Membranes with higher resistances measured in the low current density tests required generally higher applied potentials during electrolysis 251 although this overall trend was not statistically significant ($R^2 = 0.38$, p = 0.44). (Figure 3b). This lack of a 252 253 correlation contrasts with trends amongst traditional ion exchange membranes, where membrane resistance 254 at low current directly correlates to applied potential in electrochemical cells [26, 40].

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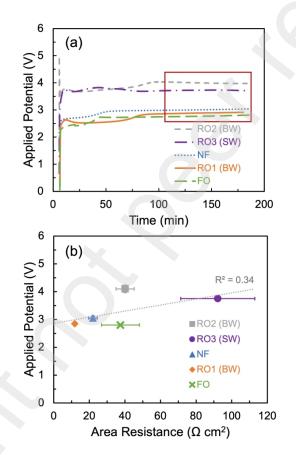


Figure 3. (a) Potential applied to electrolyze saltwater at a constant current density of 20 mA/cm² for three hours using three RO membranes, one NF membrane, and one FO membrane. The applied potential was compared when it had stabilized after two hours, indicated by the red box. (b) Comparison of the applied potentials for saltwater electrolysis and the membrane resistances. The equation of the line is y = 0.01x + 2.87.

- The generation of protons at the anode and hydroxide ions at the cathode during water electrolyzer tests produced large pH gradients between the electrolytes [12]. The final pHs after three hours of electrolysis
- 258 (Supplementary Information, Figure S7) were 1.4 ± 0.1 for the analyte and 12.2 ± 0.1 for the catholyte. RO
- and NF membranes can tolerate this pH range; however, the pH operating range of FO membranes is 3-7

[45, 48, 49]. Contact with the acidic anolyte and basic catholyte may have damaged the membrane, resultingin a lower applied potential being required for this membrane [48, 50].

262

263 *3.4 Membrane ion crossover*

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265 In resistance tests at neutral pH (at low current densities), the concentrations of H^+ and OH^- are very low 266 so charge is balanced by salt ions transported through the RO membrane. However, during water 267 electrolysis large pH gradients develop, so water ion migration plays a larger role in current transport across 268 the membrane. The difference between salt ion flux and proton or hydroxide ion flux is more significant in 269 RO, NF, and FO membranes than in CEMs because of their steric partitioning in addition to dielectric and 270 Donnan partitioning [51, 52]. While it is possible to correlate neutral membrane resistance to applied 271 potential during electrolysis for CEMs, it was not possible to correlate these two properties for RO, NF, 272 and FO membranes.

273

274 RO, NF, and FO membranes hinder proton and hydroxide transport differently than they hinder Na⁺ and 275 Cl⁻ transport because of their dense, size selective active layers with pore sizes between 0.1-2 nm. 276 Electrochemical impedance spectroscopy (EIS) was reported to measure the conductance of the active layer 277 of RO membranes in pH 3.7 and 5.7 salt solutions (both KCl and MgCl) using a rotating disk electrode 278 [27]. The active layer had a higher conductance in the solution with a higher concentration of protons (pH 279 3.7). The EIS method from this one in that it considers only the active layer of the membrane, but similar 280 trends are observed when comparing the total membrane resistance in neutral solution to applied potential during electrolysis with large pH gradients. Protons can more easily transport and carry current across the 281 282 dense polyamide layer, so increasing their concentration during electrolysis changes the membrane 283 resistance to charge transfer in comparison to the neutral salt solution.

284

285 To demonstrate the relative importance of salt ion versus water ion transport in water electrolyzer tests salt 286 ion crossover was measured using electrolytes with four different salt ions. The catholyte contained KCl 287 and the anolyte was NaClO₄. Na⁺ crossover was higher for the FO (49.5 \pm 0.5 mM) and NF (27.3 \pm 0.04 mM) membrane than for the RO membranes (RO1 = 4.4 ± 0.5 mM, RO2 = 5.3 ± 0.5 mM, RO3 = 9.8 ± 0.7 288 mM) (Figure 4). Conversely, Cl⁻ crossover was higher for the RO membranes than for the NF and FO 289 290 membranes. For Cl⁻ crossover, RO1 (BW) had 23.0 ± 0.7 mM, RO2 (BW) had 21.0 ± 0.2 mM, and RO3 291 (SW) had 10.5 ± 0.2 mM. For the NF and FO membranes Cl⁻ crossover was about four times less, with NF 292 having 6.5 ± 0.01 mM and FO having 5.5 ± 0.6 mM. K⁺ and ClO₄⁻ crossover was one to two orders of 293 magnitude less than Na⁺ and ClO_4^- crossover. The FO membrane had the highest amount of K⁺ crossover 294 $(7.5 \pm 0.4 \text{ mM})$, and the K⁺ crossover for RO2, and RO3 was undetectable. The ClO₄⁻ crossover followed 295 a similar trend as the Na⁺ crossover, with the FO and NF membranes having the most crossover (FO = 7.3 296 ± 0.05 mM, NF = 4.3 ± 0.10 mM), and the RO membranes having the least crossover (RO1 = 0.47 ± 0.07 297 mM, RO2 = 0.36 ± 0.01 , RO3 = 0.44 ± 0.01).

298 Na⁺ and Cl⁻ ions are transported across the membrane in the same direction as the electric field, and so their 299 transport is due to both diffusional and electromigration forces. Thus, these two ions had the highest amount 300 of crossover for all membranes. RO membranes showed higher Cl⁻ crossover than Na⁺, while the FO and 301 NF membranes had more Na⁺ crossover than Cl⁻. The RO and NF membranes had their active layers facing 302 the anolyte during the experiments. The proton generation at the anode caused an anolyte pH between 1-2 303 for all the membranes, (Supplementary Information, Figure S8). According to published zeta potential data of RO membrane active layers, the active layer of the membrane is most likely positively charged at a pH 304 305 between 1-2 [28]. During electrolysis, this positively charged active layer in contact with the acidic anolyte 306 could have caused the preferential transport of Cl⁻ over Na⁺ in the direction of the potential gradient [28]. 307 Hydroxide generation at the cathode resulted in a pH between 12-13 in the catholyte (Supplementary 308 Information, Figure S8). By changing the active layer to face the catholyte instead of the anolyte, the 309 membrane charge will most likely be negative instead, and Cl⁻ transport could be further reduced in comparison to Na⁺ transport, as suggested by results in a previous study where the impact of the direction of active layer was examined on salt ion transport [12]. There was minimal K⁺ and ClO_4^- crossover here for all membranes because the concentration and potential gradients for these ions were in directions opposite to the electric field.

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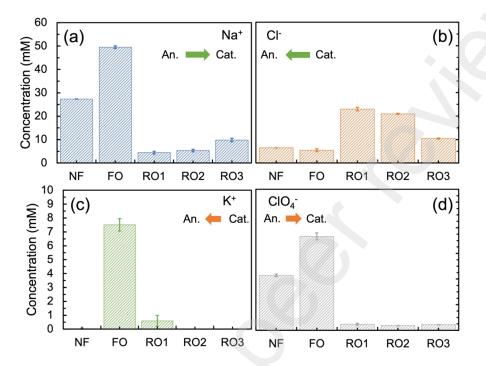


Figure 4. Concentration of salt ions that crossed over the membrane after a current density of 20 mA/cm2 was applied to the cylindrical reactor for one hour with 1 M NaClO₄ anolyte and 1 M KCl catholyte. Concentrations of ions that crossed over into the opposite chamber (a) sodium, (b) chloride, (c) potassium ions, and (d) perchlorate. Green arrows indicate ions with concentration and potential gradients in the same direction, while orange arrows indicate ions with concentration and potential gradients in opposite directions. RO 1 and 2 are BW membranes, RO 3 is an SW membrane.

315 Using the ion crossover data and the total Coulombs of charge passed through the electrolyzer based on the

- set current, we calculated the fraction of charge carrier that was due to water ions (protons and hydroxide
- 317 ions) for each membrane (calculations in the Supplementary Information). The membranes with the lowest

fraction of charge carrier (FO and NF) across the membrane had the highest amount of salt transport in the

direction of the potential gradient which balanced charge across the membrane (Figure 5a). The RO

320 membranes hindered salt ion transport during electrolysis more so than the NF and CTA FO membranes.

- The FO membrane had the highest amount of total salt crossover, and RO3 (SW) had the least amount of
- total salt crossover and highest fraction of proton and hydroxide charge carriers.
- 323

The FO membrane has a more homogenous active layer with higher free volume, making it a looser membrane, so salt ion transport across the membrane could rapidly occur (even during the brief few minutes when the chronopotentiometry experiment was being set up and initial salt samples were being collected), causing this membrane to have a fraction of charge carrier for protons and hydroxide close to zero [53]. Also, the FO membrane may have been damaged during electrolysis due to its smaller tolerance for a very high or low pH, which could have allowed salt ions to easily move between compartments during sample collection after the experiment. The smaller pore sizes and higher salt rejections of RO membranes may

331 have contributed to them hindering ion transport more effectively than the NF and FO membranes [37].

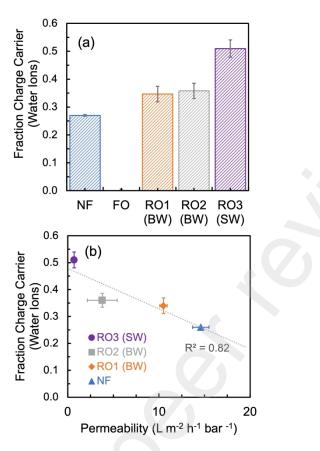


Figure 5. (a) The fraction of charge that was carried by protons and hydroxide ions for each membrane during the ion crossover experiment. (b) Comparison of the fraction of charge carriers that was protons and hydroxide to the membrane permeability measured using the high-pressure dead-end cell. The equation of the line is y = -0.01x + 0.48.

332 The RO3 (SW) membrane has the highest reported salt rejection according to manufacturer spec sheets and

- the highest measured membrane resistance to Na⁺ and Cl⁻ transport, indicating the membrane sufficiently
 hindered the salt transport and preferentially transported the electrochemically active species (protons and
 hydroxide ions).
- 336

337 The fraction charge carrier of protons for each membrane had an inverse trend with membrane permeability. 338 The RO3 (SW) membrane had the lowest permeability and had the highest fraction of proton and hydroxide 339 as charge carriers, while the membrane with the highest permeability (NF) had the lowest fraction of proton 340 and hydroxide as charge carriers (Figure 5b). This trend indicated that the membranes that most effectively hindered salt ion transport resulted in the charge being balanced by increased proton or hydroxide transport, 341 although likely this trend was not significant ($R^2 = 0.82$, p = 0.13), which may be due to the small number 342 343 of data (four membranes). Additional tests will be needed to further examine the significance of this 344 apparent trend. 345

346 4. Conclusions

347

An evaluation of the electrochemical properties of four types of asymmetric filtration membranes shows that RO, NF, and FO membrane resistance during electrolysis cannot be predicted using tests developed for ion exchange membranes due to the large pH gradients that will develop in water electrolyzer tests. Results from the ion crossover tests show a general trend of increased charge balance based on water ions rather than salt ions during electrolysis. Charge is balanced by salt ions in resistance tests at neutral pH, while
 charge is balanced by both salt ions and water ions in water electrolyzer tests. As a result of this higher
 dependence of water ions in the water electrolyzer tests, there was no correlation between resistances
 measured at low current densities with overpotentials measured in water electrolyzer tests.

357 Acknowledgements:

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362 363

364 References

- 365 [1] S. Dresp, F. Dionigi, M. Klingenhof, and P. Strasser, Direct electrolytic splitting of seawater: Opportunities and challenges, Acs Energy Lett. 4 (2019) 933-942.
 367 https://doi.org/10.1021/acsenergylett.9b00220.
- B.E. Logan, L. Shi, and R. Rossi, Enabling the use of seawater for hydrogen gas production
 in water electrolyzers, Joule 5 (2021) 760-762.
 https://doi.org/10.1016/j.joule.2021.03.018.
- W.M. Tong, M. Forster, F. Dionigi, S. Dresp, R.S. Erami, P. Strasser, A.J. Cowan, and P.
 Farras, Electrolysis of low-grade and saline surface water, Nat. Energy 5 (2020) 367-377.
 https://doi.org/10.1038/s41560-020-0550-8.
- K. Meier, Hydrogen production with sea water electrolysis using Norwegian offshore wind
 energy potentials, Int. J. Energy Environ. Eng. 5 (2014) <u>https://doi.org/10.1007/s40095-014-0104-6</u>.
- R. d'Amore-Domenech, O. Santiago, and T.J. Leo, Multicriteria analysis of seawater
 electrolysis technologies for green hydrogen production at sea, Renew. Sust. Energy Rev.
 133 (2020) https://doi.org/10.1016/j.rser.2020.110166.
- Y. Yang, J. Shin, J.T. Jasper, and M.R. Hoffmann, Multilayer heterojunction anodes for saline wastewater treatment: Design strategies and reactive species generation mechanisms, Environ. Sci. Technol. 50 (2016) 8780-8787.
 <u>https://doi.org/10.1021/acs.est.6b00688</u>.
- E. Asghari, M.I. Abdullah, F. Foroughi, J.J. Lamb, and B.G. Pollet, Advances, opportunities, and challenges of hydrogen and oxygen production from seawater electrolysis: An electrocatalysis perspective, Curr. Opin. Electroche. 31 (2022) https://doi.org/10.1016/j.coelec.2021.100879.
- F. Dionigi, T. Reier, Z. Pawolek, M. Gliech, and P. Strasser, Design criteria, operating conditions, and nickel-iron hydroxide catalyst materials for selective seawater electrolysis, Chemsuschem 9 (2016) 962-972. https://doi.org/10.1002/cssc.201501581.
- S. Dresp, F. Dionigi, S. Loos, J.F. de Araujo, C. Spori, M. Gliech, H. Dau, and P. Strasser,
 Direct electrolytic splitting of seawater: Activity, selectivity, degradation, and recovery
 studied from the molecular catalyst structure to the electrolyzer cell level, Adv. Energy
 Mater. 8 (2018) <u>https://doi.org/10.1002/aenm.201800338</u>.
- S. Dresp, T.N. Thanh, M. Klingenhof, S. Bruckner, P. Hauke, and P. Strasser, Efficient direct seawater electrolysers using selective alkaline NiFe-LDH as OER catalyst in asymmetric electrolyte feeds, Energy Environ. Sci. 13 (2020) 1725-1729.
 <u>https://doi.org/10.1039/d0ee01125h</u>.
- J.S. Ko, J.K. Johnson, P.I. Johnson, and Z.Y. Xia, Decoupling oxygen and chlorine
 evolution reactions in seawater using iridium-based electrocatalysts, Chemcatchem 12
 (2020) 4526-4532. https://doi.org/10.1002/cctc.202000653.
- 402 [12] L. Shi, R. Rossi, M. Son, D.M. Hall, M.A. Hickner, C.A. Gorski, and B.E. Logan, Using reverse osmosis membranes to control ion transport during water electrolysis, Energy Environ. Sci. 13 (2020) 3138-3148. <u>https://doi.org/10.1039/d0ee02173c</u>.
- P.M. Biesheuvel, L. Zhang, P. Gasquet, B. Blankert, M. Elimelech, and W.G.J. van der Meer, Ion selectivity in brackish water desalination by reverse osmosis: Theory, measurements, and implications, Environ. Sci. Tech. Let. 7 (2020) 42-47.
 https://doi.org/10.1021/acs.estlett.9b00686.

- 409[14]V. Freger and G.Z. Ramon, Polyamide desalination membranes: Formation, structure, and410properties,Prog.Polym.Sci.122(2021)411https://doi.org/10.1016/j.progpolymsci.2021.101451.
- [15] N. Fridman-Bishop and V. Freger, What makes aromatic polyamide membranes superior:
 New insights into ion transport and membrane structure, J. Membr. Sci. 540 (2017) 120128. https://doi.org/10.1016/j.memsci.2017.06.035.
- 415 [16] M. Hafiz, A.H. Hawari, R. Alfahel, M.K. Hassan, and A. Altaee, Comparison of 416 nanofiltration with reverse osmosis in reclaiming tertiary treated municipal wastewater for 417 irrigation purposes, Membranes-Basel 11 (2021) 418 https://doi.org/10.3390/membranes11010032.
- 419 [17] J. Benavente and C. Fernandezpineda, Electrokinetic phenomena in porous membranes 420 Determination of phenomenological coefficients and transport numbers, J. Membr. Sci. 23
 421 (1985) 121-136. https://doi.org/10.1016/S0376-7388(00)82214-0.
- 422 [18] G. Jonsson and J. Benavente, Determination of some transport-coefficients for the skin and
 423 porous layer of a composite membrane, J. Membr. Sci. 69 (1992) 29-42.
 424 <u>https://doi.org/10.1016/0376-7388(92)80165-G</u>.
- J. Benavente and G. Jonsson, Electrokinetic characterization of composite membranes:
 Estimation of different electrical contributions in pressure induced potential measured
 across reverse osmosis membranes, J. Membr. Sci. 172 (2000) 189-197.
 https://doi.org/10.1016/S0376-7388(00)00325-2.
- L. Lin, C.C. Feng, R. Lopez, and O. Coronell, Identifying facile and accurate methods to measure the thickness of the active layers of thin-film composite membranes - A comparison of seven characterization techniques, J. Membr. Sci. 498 (2016) 167-179.
 https://doi.org/10.1016/j.memsci.2015.09.059.
- 433 [21] S. Habib and S.T. Weinman, A review on the synthesis of fully aromatic polyamide reverse
 434 osmosis membranes, Desalination 502 (2021)
 435 https://doi.org/10.1016/j.desal.2021.114939.
- 436 [22] R.J. Petersen, Composite reverse-osmosis and nanofiltration membranes, J. Membr. Sci.
 437 83 (1993) 81-150. <u>https://doi.org/10.1016/0376-7388(93)80014-0</u>.
- 438 [23] D.L. Shaffer, J.R. Werber, H. Jaramillo, S.H. Lin, and M. Elimelech, Forward osmosis:
 439 Where are we now?, Desalination 356 (2015) 271-284.
 440 https://doi.org/10.1016/j.desa12014.10.031.
- 441 [24] G.J. Hwang, H. Ohya, and T. Nagai, Ion exchange membrane based on block copolymers.
 442 Part III: Preparation of cation exchange membrane, J. Membr. Sci. 156 (1999) 61-65.
 443 https://doi.org/10.1016/S0376-7388(98)00331-7.
- 444 [25] A.H. Galama, N.A. Hoog, and D.R. Yntema, Method for determining ion exchange 445 membrane resistance for electrodialysis systems, Desalination 380 (2016) 1-11. 446 <u>https://doi.org/10.1016/j.desal.2015.11.018</u>.
- J. Kamcev, R. Sujanani, E.S. Jang, N. Yan, N. Moe, D.R. Paul, and B.D. Freeman, Salt concentration dependence of ionic conductivity in ion exchange membranes, J. Membr.
 Sci. 547 (2018) 123-133. https://doi.org/10.1016/j.memsci.2017.10.024.
- [27] N. Fridman-Bishop and V. Freger, When salt-rejecting polymers meet protons: An electrochemical impedance spectroscopy investigation, Langmuir 33 (2017) 1391-1397.
 https://doi.org/10.1021/acs.langmuir.6b04263.
- 453 [28] M. Stolov and V. Freger, Membrane charge weakly affects ion transport in reverse osmosis,
 454 Environ. Sci. Tech. Let. 7 (2020) 440-445. <u>https://doi.org/10.1021/acs.estlett.0c00291</u>.

- 455 [29] M. Stolov and V. Freger, Degradation of polyamide membranes exposed to chlorine: An
 456 impedance spectroscopy study, Environ. Sci. Technol. 53 (2019) 2618-2625.
 457 https://doi.org/10.1021/acs.est.8b04790.
- 458 [30] G. Tiravanti, The direct-current method for measuring charged membrane conductance, J.
 459 Membr. Sci. 9 (1981) 229-243. <u>https://doi.org/10.1016/S0376-7388(00)80266-5</u>.
- 460 [31] G.M. Geise, A.J. Curtis, M.C. Hatzell, M.A. Hickner, and B.E. Logan, Salt concentration differences alter membrane resistance in reverse electrodialysis stacks, Environ. Sci. Tech.
 462 Let. 1 (2014) 36-39. <u>https://doi.org/10.1021/ez4000719</u>.
- 463 [32] W. Oelssner, F. Berthold, and U. Guth, The iR drop well-known but often underestimated
 464 in electrochemical polarization measurements and corrosion testing, Mater Corros 57
 465 (2006) 455-466. <u>https://doi.org/10.1002/maco.200603982</u>.
- 466 [33] M. Arif, S.C.P. Cheung, and J. Andrews, A systematic approach for matching simulated
 467 and experimental polarization curves for a PEM fuel cell, Int. J. Hydrogen Energy 45
 468 (2020) 2206-2223. https://doi.org/10.1016/j.ijhydene.2019.11.057.
- 469 [34] M. Carmo, D.L. Fritz, J. Merge, and D. Stolten, A comprehensive review on PEM water
 470 electrolysis, Int. J. Hydrogen Energy 38 (2013) 4901-4934.
 471 https://doi.org/10.1016/j.ijhydene.2013.01.151.
- 472 [35] R. Phillips and C.W. Dunnill, Zero gap alkaline electrolysis cell design for renewable energy storage as hydrogen gas, Rsc Adv 6 (2016) 100643-100651.
 474 https://doi.org/10.1039/c6ra22242k.
- 475 [36] Y. Okamoto and J.H. Lienhard, How RO membrane permeability and other performance
 476 factors affect process cost and energy use: A review, Desalination 470 (2019)
 477 https://doi.org/10.1016/j.desal.2019.07.004.
- [37] R. Epsztein, R.M. DuChanois, C.L. Ritt, A. Noy, and M. Elimelech, Towards single-species selectivity of membranes with subnanometre pores, Nat Nanotechnol 15 (2020)
 480 426-436. <u>https://doi.org/10.1038/s41565-020-0713-6</u>.
- 481 [38] C. Boo, Y.K. Wang, I. Zucker, Y. Choo, C.O. Osuji, and M. Elimelech, High performance nanofiltration membrane for effective removal of perfluoroalkyl substances at high water recovery, Environ. Sci. Technol. 52 (2018) 7279-7288.
 484 https://doi.org/10.1021/acs.est.8b01040.
- 485 [39] "TriSep ACM5 Low Energy RO Membrane," in "Product Specification Trisep ACM5,"
 486 TriSep, 2021.
- 487 [40] K.R. Cooper and M. Smith, Electrical test methods for on-line fuel cell ohmic resistance
 488 measurement, J. Power Sources 160 (2006) 1088-1095.
 489 https://doi.org/10.1016/j.jpowsour.2006.02.086.
- 490 [41] J.C. Diaz and J. Kamcev, Ionic conductivity of ion-exchange membranes: Measurement
 491 techniques and salt concentration dependence, J. Membr. Sci. 618 (2021)
 492 https://doi.org/10.1016/j.memsci.2020.118718.
- A. Ramdani, A. Deratani, S. Taleb, N. Drouiche, and H. Lounici, Performance of NF90 493 [42] 494 and NF270 commercial nanofiltration membranes in the defluoridation of Algerian brackish Desalin Water 495 water. Treat. 212 (2021)286-296. 496 https://doi.org/10.5004/dwt.2021.26680.
- 497 [43] M. Manttari, T. Pekuri, and M. Nystrom, NF270, a new membrane having promising characteristics and being suitable for treatment of dilute effluents from the paper industry,
 499 J. Membr. Sci. 242 (2004) 107-116. https://doi.org/10.1016/j.memsci.2003.08.032.

- 500 [44] K. Kosutic, L. Kastelan-Kunst, and B. Kunst, Porosity of some commercial reverse osmosis and nanofiltration polyamide thin-film composite membranes, J. Membr. Sci. 168 (2000) 101-108. <u>https://doi.org/10.1016/S0376-7388(99)00309-9</u>.
- 503[45]"Product data sheet FilmTec NF270 nanofiltration elements for commercial systems,"504DuPont,2021.[Online].Available:505https://www.dupont.com/content/dam/dupont/amer/us/en/water-solutions/public/documents/en/NF-FilmTec-NF270-PDS-45-D01529-en.pdf
- [46] Z.Y. Wang, Z.X. Wang, S.H. Lin, H.L. Jin, S.J. Gao, Y.Z. Zhu, and J. Jin, Nanoparticle-templated nanofiltration membranes for ultrahigh performance desalination, Nat Commun
 9 (2018) https://doi.org/10.1038/s41467-018-04467-3.
- 510 [47] L.F. Greenlee, D.F. Lawler, B.D. Freeman, B. Marrot, and P. Moulin, Reverse osmosis
 511 desalination: Water sources, technology, and today's challenges, Water Res. 43 (2009)
 512 2317-2348. <u>https://doi.org/10.1016/j.watres.2009.03.010</u>.
- 513[48]"OsmoF2OTM Forward Osmosis Membranes Flat Sheet Membranes," Fluid Technology514Solutions,Inc.,2018.[Online].Available:515https://www.sterlitech.com/media/wysiwyg/FTSDataSheet.pdf
- 516 [49] DuPont, "FilmTech XLE-440 Element," in "Product Data Sheet," 2020. [Online].
 517 Available: <u>https://www.dupont.com/content/dam/dupont/amer/us/en/water-</u>
 518 solutions/public/documents/en/RO-FilmTec-XLE-440-PDS-45-D01513-en.pdf
- 519 [50] G. Li, X.M. Li, T. He, B. Jiang, and C.J. Gao, Cellulose triacetate forward osmosis
 520 membranes: preparation and characterization, Desalin Water Treat. 51 (2013) 2656-2665.
 521 <u>https://doi.org/10.1080/19443994.2012.749246</u>.
- L. Wang, T.C. Cao, J.E. Dykstra, S. Porada, P.M. Biesheuvel, and M. Elimelech, Salt and water transport in reverse osmosis membranes: Beyond the solution-diffusion model, Environ. Sci. Technol. 55 (2021) 16665-16675. https://doi.org/10.1021/acs.est.1c05649.
- [52] R. Epsztein, E. Shaulsky, N. Dizge, D.M. Warsinger, and M. Elimelech, Role of ionic charge density in Donnan exclusion of monovalent anions by nanofiltration, Environ. Sci.
 527 Technol. 52 (2018) 4108-4116. https://doi.org/10.1021/acs.est.7b06400.
- 528 [53] S.J. Kim, S. Kook, B.E. O'Rourke, J. Lee, M. Hwang, Y. Kobayashi, R. Suzuki, and I.S.
 529 Kim, Characterization of pore size distribution (PSD) in cellulose triacetate (CTA) and
 530 polyamide (PA) thin active layers by positron annihilation lifetime spectroscopy (PALS)
 531 and fractional rejection (FR) method, J. Membr. Sci. 527 (2017) 143-151.
 532 <u>https://doi.org/10.1016/j.memsci.2016.12.064</u>.

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